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1 **Return flows from beaver ponds enhance floodplain-to-river metals**
2 **exchange in**
3 **alluvial mountain catchments**

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24Abstract

25River to floodplain hydrologic connectivity is strongly enhanced by beaver-
26(*Castor canadensis*) engineered channel water diversions. The
27hydroecological impacts are wide ranging and generally positive, however,
28the hydrogeochemical characteristics of beaver-induced flowpaths have not
29been thoroughly examined. Using a suite of complementary ground- and
30drone-based heat tracing and remote sensing methodology we characterized
31the physical template of beaver-induced floodplain exchange for two alluvial
32mountain streams near Crested Butte, Colorado, USA. A flowpath-oriented
33perspective to water quality sampling allowed characterization of the
34chemical evolution of channel water diverted through floodplain beaver
35ponds and ultimately back to the channel in 'beaver pond return flows'.
36Return seepages were universally suboxic, while ponds and surface return
37flows showed a range of oxygen concentration due to in-situ photosynthesis
38and atmospheric mixing. Median concentrations of reduced metals:
39manganese (Mn), iron (Fe), aluminum (Al), and arsenic (As) were
40substantially higher along beaver-induced flowpaths than in geologically
41controlled seepages and upstream main channel locations. The areal
42footprint of reduced return flow seepage flowpaths were imaged with surface
43electromagnetic methods, indicating extensive zones of high-conductivity

44shallow groundwater flowing back toward the main channels and emerging
45at relatively warm bank seepage zones observed with infrared. Multiple-
46depth redox dynamics within one focused seepage zones showed coupled
47variation over time, likely driven by observed changes in seepage rate that
48may be driven by pond stage. High-resolution times series of dissolved Mn
49and Fe collected downstream of the beaver-impacted reaches indicated
50seasonal dynamics in mixed river metal concentrations. Al time series
51concentrations showed proportional change to Fe at the smaller stream
52location, indicating chemically reduced flowpaths were sourcing Al to the
53channel. Overall our results indicated beaver-induced floodplain exchanges
54create important, and perhaps dominant, transport pathways for floodplain
55metals by expanding chemically-reduced zones paired with strong advective
56exchange.

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59**Key words:** river; groundwater/surface water interactions; beaver;

60floodplain; drone; water quality

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621. Introduction

63 The concept of 'river corridor' science recognizes that the quality of
64 flowing surface waters is intrinsically linked to their contributing catchments
65 through hydrologic connectivity, including lower terrestrial hillslopes,
66 floodplains, and riparian zones (Covino, 2017; Poole, 2010; Vidon et al.,
67 2010). Bidirectional river-floodplain exchange in particular can be critical to
68 basin water storage and nutrient transformation dynamics (Harvey and
69 Gooseff, 2015), yet floodplain hydrologic exchange flows are often driven
70 primarily by episodic high-flow events (e.g. Sawyer et al., 2014) or relatively
71 slow-exchanging, long hyporheic flowpaths (Boano et al., 2014). Beaver
72 (*Castor canadensis*) disrupt these abiotic floodplain exchange drivers by
73 actively diverting large quantities of channel water laterally using an
74 engineered series of dams, impacting both wet and dry season floodplain
75 connection (Westbrook et al., 2006). As humans allow beaver to return to
76 their extensive natural habitats across North America, the fundamental
77 dynamics of river corridor hydrologic connectivity are being strongly altered
78 toward a template of spatial 'discontinuum' and enhanced exchange
79 (Burchsted et al., 2010). Much beaver-induced floodplain disturbance is
80 undoubtedly viewed as a net positive in the context of natural and efficient
81 watershed restoration. But as beaver ponds and seepage zones
82 accompanying beaver activity often exhibit suboxic to hypoxic conditions
83 (Collen and Gibson, 2000), there is the potential to mobilize large quantities

84of reduced metals and associated contaminants from alluvial sediments to
85streams and rivers.

86 Beaver populations are steadily increasing in North America due to
87stricter trapping laws, a general decline in trapping interest, passive and
88active conservation efforts, and a relative absence of predators (Hood,
892011). In one sense, this rebound can be viewed as North American
90watersheds returning to a natural, or pre-European settlement state one that
91is extensively engineered by the beaver. Before European settlement the
92North American beaver population numbered between 60-400 million
93individuals (Seton, 1929), and their dams and foraging influenced almost
94every floodplain system from arid Mexico to the arctic tundra (Naiman et al.,
951988). After the arrival of significant numbers of Europeans in the early
961600's beaver populations declined in response to extreme trapping until the
97animal was functionally extinct on the continent by 1900. Before trapping
98began, most rivers in North America had extensive beaver-induced
99floodplains and numerous wood snags that retained carbon and nutrients in
100the headwaters. Evidence of the extensive effects on the riparian landscape
101of large historical populations can still be seen hundreds of years later
102(Naiman et al., 1986).

103 Beaver construct dams across streams and wetlands to increase
104habitat favorable to their basic needs of forage and protection.
105Impoundments can have a variety of influences on the physical and

106biological characteristics of floodplain areas and riparian zones. Analogous to
107anthropogenic dams, the combined effects of beaver dams is often a
108reduction in peak river discharge, a smoothing of catchment outlet
109hydrograph (Ligon et al., 1995), and a general increase in reach-scale water
110residence time (Jin et al., 2009). These moderated flow patterns can stabilize
111the stream channel and decrease bed mobilization by reducing erosive
112forces, such as shear stress on the stream bank and along the sediment-
113water interface. Storage of water behind the dams can be an appreciable
114fraction of the catchment surficial water budget during dry periods and often
115results in greater connection between riparian vegetation and the water
116table throughout the year. An analysis of aerial photo mosaics from 1948 to
1172002 from Central Alberta, Canada indicated that the number of 'active'
118beaver lodges could explain greater than 80% of the variability in floodplain
119open water area through a number of wet and dry periods (Hood and Bayley,
1202008). However, the beaver-induced water storage story is complex. Some
121studies have also suggested that the ecological effects of increased water
122storage may be negated in part by amplification of the evaporative flux due
123to an increase in stream and floodplain surface water area (Collen and
124Gibson, 2000). Recent work has indicated that beaver-induced recharge of
125alluvial floodplains may not substantially increase late summer low-flows, in
126part because much of this floodplain water may be trapped in low
127permeability soils and/or is simply recently diverted channel water (Nash et
128al., 2018). However, flows returning to the channel from reactive beaver-

129 induced floodplain storage are likely to be important conduits of carbon
130 transport (Catalán et al., 2017) and nutrient transformation (Briggs et al.,
131 2013; Wegener et al., 2017) throughout the year.

132 The recent phenomena of encouraging beaver recolonization and the
133 installation of anthropogenic ‘beaver dam analogues’ in the context of
134 stream restoration hopes to capitalize on expected net positive
135 hydrogeological and ecosystem impacts (Lautz et al., 2018; Pilliod et al.,
136 2018; Wohl et al., 2015). Natural and simulated dams can mitigate incised
137 western channels by increasing floodplain connection and riparian vegetation
138 regrowth, which in turn positively influences desirable recreational fish
139 populations (Bouwes et al., 2016). The channel water temperature response
140 to beaver-induced floodplain connection is complicated and spatially
141 heterogeneous (Majerova et al., 2015), but it has been shown to moderate
142 the warmest daily temperatures and provide cool thermal refugia for
143 stressed aquatic species in Oregon (Weber et al., 2017). Benefits of beaver
144 colonization that may outweigh complications to infrastructure have even
145 been recently recognized for urban drainages (Bailey et al., 2019). However,
146 not all impacts of natural and simulated beaver dams will be desirable to
147 humans (Pilliod et al., 2018). Negative impacts of result mainly from the
148 raising of the water table adjacent to the stream, flooding, impoundment of
149 drainage systems, and the cutting of desirable vegetation (Collen and
150 Gibson, 2000).

151 Gradients in pH and redox conditions along chemical pathways
152 between riparian soils and stream sediments are also strongly affected by
153 beaver impoundment. For example, in some systems oxic soils move/regress
154 farther away from the original channel as the water level rises, and the
155 magnitude of anoxic soil area grows accordingly (Naiman et al., 1988).
156 Anoxic conditions can readily develop on the floodplain due to the tranquil
157 flow regime of beaver ponds and large supply of local organic carbon, and
158 along biogeochemically reactive subsurface flowpaths that route floodplain
159 water back to the channel, creating 'natural reduced zones' (NRZs, e.g.
160 Dwivedi et al., 2018). Anoxic soils may increase the acid neutralizing
161 capacity of the soil water due to the retention of nitrate and sulfate, and act
162 as a net source of iron and ammonium ions (Cirimo and Driscoll, 1993),
163 though quantitative research regarding beaver-induced mobilization of
164 reduced chemical species is generally lacking. Microbial activity of
165 floodplains and riparian zones has been found to be greatly increased by
166 impoundments, a process that has important implications from the pore- to
167 the landscape-scale (Wegener et al., 2017). Enhanced biogeochemical
168 reactivity and potential mobilization of floodplain nutrients, metals, and
169 contaminants necessitates a more complete understanding as beavers are
170 increasingly regarded as a stream restoration solution. Recent studies have
171 signaled beaver-related stream restoration practices may be outpacing
172 fundamental science regarding the wide ranging physical and chemical
173 impacts of such projects (Lautz et al., 2018; Pilliod et al., 2018).

174 For river corridor hydrologic exchanges to be influential to water
175quality, hydrologic exchange flows need to be both chemically reactive and
176of appreciable volume compared to river discharge so as to alter net channel
177solute transport dynamics (Wondzell, 2011). While biogeochemically reactive
178cross-meander bend hyporheic exchange may be prevalent in most alluvial
179river corridors, a combination of relatively low hydraulic gradient and tight
180floodplain soils can limit the impact of this exchange on net river chemistry
181(Pai et al., 2017), particularly at the km-reach scale. In contrast, beaver
182dams are known to push large volumes of surface water laterally into the
183floodplain through engineered fill and spill pathways. Here, we characterize
184the natural metal transformation and transport dynamics of a specific type of
185river corridor hydrologic exchange flow, termed here: ‘beaver pond return
186flows.’ Like the more well-known ‘irrigation return flows’ to rivers driven by
187the application of water to adjacent cropland (Essaid and Caldwell, 2017),
188beaver pond return flows are enabled by the purposeful redirection of water
189outside of the channel. Redirected channel water that is not lost to floodplain
190evapotranspiration returns to the river in a spectrum of surface flows and
191subsurface seepage zones (Majerova et al., 2015). By using a combination of
192remote sensing and direct contact measurements, we identify beaver-
193induced floodplain exchange flowpaths along two two alluvial mountain
194streams of varied size. Chemical measurements collected across a full-year
195hydrological cycle (neglecting winter months) at the pond return flows, and
196at other observed groundwater seepage types, indicate beaver-induced

197floodplain exchange can be a dominant mechanism of natural metals flux
198along alluvial river corridors. Further, extensive deposition of solid metal
199oxides at return flow discharge points likely provides an important
200source/sink function for a variety of contaminant transport problems. The
201evidence to support these statements is shown in the following sections.

2022. **Materials and Methods**

203 A suite of complementary ground- and drone-based heat tracing and
204remote sensing methodology was used to characterize the hydrogeological
205template of beaver-induced floodplain exchange for two alluvial mountain
206streams of disparate size. A flowpath-oriented perspective to water quality
207sampling allowed characterization of the chemical evolution of channel water
208diverted through floodplain ponds and ultimately returned to the channel.

2092.1 *Study Area*

210 The Lawrence Berkeley National Laboratory Watershed Function
211Scientific Focus Area (SFA) has established an experimental watershed
212encompassing the drainages of the East River near Crested Butte, Colorado
213(USA) to quantify the myriad nested processes impacting the ability of
214mountainous systems to retain and release water, nutrients, carbon, and
215metals. This scientific ‘community watershed’ hosts ongoing research
216spanning a wide range of spatial scales and physical, chemical, and
217biological processes. The SFA encompasses the drainages of the East River,
218Washington Gulch, Slate River (including Oh-Be-Joyful Creek), and Coal Creek

219(Figure 1a). Although each watershed has analogous sections of meandering
220alluvial stream, they also display unique flow dynamics, current land use
221practices, and legacies of mining-related contamination. While there is some
222direct impact from cattle ranching along the East River corridor, that system
223is generally considered a 'pristine' end-member due to the lack of substantial
224mining activity and ore-rich rock draining its environs. In contrast, Coal Creek
225and the Slate River are more highly influenced by heavy metals, such as
226arsenic, copper, cadmium, and zinc due to legacy mine activities in those
227drainages. For the focus areas of the current study, we chose analogous,
228meandering open valley sections of the larger East River and smaller Coal
229Creek with observed contemporary beaver inhabitation (Figure 1a). This
230research began as broader investigation of natural metal mobility and metal
231oxide deposition at geologically-controlled groundwater seepages throughout
232the SFA; however, early in the study, it became apparent that beaver pond
233return flows were likely to be an important metals flux pathway to consider,
234and the research plan was adapted accordingly.

2352.2 *Aerial mapping of floodplain beaver ponds*

236 Floodplain zones inundated with beaver-induced exchange of channel
237water are difficult to navigate on the ground, but the typical open canopy
238nature of such areas presents opportunity for small unoccupied aerial vehicle
239(sUAS, or 'drone') and satellite-based mapping techniques. We deployed
240various multicopter sUAS (3D Robotics Solo, 3D Robotics, Berkeley, CA) at the
241East River reach from August 12-17, 2017 and July 28-August 2, 2018; and at

242Coal Creek on July, 31 2018. During sequential flights the sUAS were
243equipped with various sensors, similar to the approach described by Briggs
244et al., (2018). For high-resolution visible imagery we used a Ricoh GRII
245Camera (Ricoh Imaging Company, Ltd., Japan). Image stills from multiple
246flight lines, altitudes (generally 200-350 ft above ground surface), and
247directions were compiled into larger “stitched,” georectified orthoimages of
248the river corridor using Agisoft PhotoScan software. Position of the aircraft
249was tracked by internal GPS, and although ground control points were
250deployed for some flights, they were not used in postprocessing of the visible
251imagery. Structure from motion (SfM) techniques were then applied using
252Agisoft PhotoScan software to generate time-specific surface digital
253elevation models of floodplain structure and exposed channel
254geomorphology. Details regarding the UAS sensor specifications and the
255calculated spatial precision of the compiled orthoimages are listed in the
256public data release of this data at Briggs et al., (2019b). Although the
257imagery from satellites is of substantially lower resolution than that
258achievable with sUAS, there may be an existing wealth of historical imagery
259available to assess longer term beaver pond structure dynamics. We used
260Google Earth (Google, Mountain View, CA, USA) to qualitatively assess
261beaver occupation of the Coal Creek reach back to 1999 (earliest available
262clear imagery).

2632.3 *Geolocation and characterization of seepage zones*

264 Thermal infrared is sensitive to the water surface 'skin' temperature
265 (Handcock et al., 2006) and can be used to geolocate river corridor seepage
266 zones and identify surface water flow patterns at times of natural thermal
267 contrast (Dugdale, 2016; Hare et al., 2015). We expected discharge of
268 deeper groundwater to approximate the surface annual mean temperature
269 (approximately 8°C, Constantz, 2008) whereas shallow groundwater
270 discharge and pond return flows should be warmer in summer, providing
271 multiple characteristic targets for infrared imaging. Thermal infrared data
272 were collected on the ground using handheld FLIR i7 and T600bx series
273 cameras (FLIR Systems, Wilsonville, OR) throughout the beaver-impacted
274 reaches and along an additional approximate 6 km of the upper East River
275 and within the nearby Oh-Be-Joyful Creek drainage. The purpose of the
276 larger-scale thermal infrared mapping was to identify a range of dominant
277 non-beaver-impacted groundwater discharge (seepage) zones for
278 geochemical characterization. To augment the ground-based thermal
279 surveys throughout the beaver-impacted floodplain areas we collected
280 radiometric thermal infrared data from sUAS using a gimbal-mounted FLIR
281 VUE Pro R 13mm camera.

282 Because thermal infrared imaging may not reliably locate submerged
283 seepage zones, particularly in fast flowing rivers (Hare et al., 2015), armored
284 fiber-optic distributed temperature sensing (FO-DTS) cables were deployed
285 along an approximate 2.4 km floodplain channel length at the East River

286from August 15 to August 22, 2017. FO-DTS technology for environmental
287temperature sensing is thoroughly reviewed by Tyler et al., (2009). Effort
288was made to emplace the weighted cables along the sediment-water
289interface of the 'cutting' banks of meander bends as these locations typically
290show enhanced exchange of surface and groundwater. FO-DTS data were
291collected with a Sensonet Oryx control unit (Sensonet Ltd., United
292Kingdom) run in double-ended mode at 10-min acquisition time per channel
293(20-min per measurement) and 1.01 m linear spatial resolution.

294 Once seepage zones of various type were identified, stream flow was
295physically gauged for several of the higher volume discharges using small
296custom surface weirs, graduated cylinders, and a stopwatch. Slow flowing
297'diffuse' seepage rates were evaluated at 4 discrete locations along an East
298River side channel margin, down gradient of a large beaver pond, where
299seepage was indicated by thermal imaging and Fe-oxide staining from
300August 23 to November 4, 2017. For comparison, seepage was also
301monitored over this period within an adjacent spatially focused, higher flow
302beaver pond return seepage zone. Vertical seepage rates were tracked over
303time using profiles of shallow (0.01, 0.07, 0.11 m depth below sediment-
304water interface) saturated sediment temperatures collected with iButton
305thermal data loggers (Maxim Integrated DS1922L) run at 0.0625 °C precision
306embedded in short steel pipes, as described in detail by Briggs et al. (2014).
307The workflow suggested by Irvine et al., (2017) that combines diurnal
308temperature signal-based thermal parameter measurements with diurnal

309signal amplitude attenuation was used to perform the analytical modeling of
310vertical water flux rates. The fundamental (diurnal) sinusoids were derived
311from the raw temperature data using the Captain Toolbox (Young et al.,
3122010) and VFLUX2 (Irvine et al., 2015) Matlab-based programs. Vertical flux
313was evaluated over time with the amplitude ratio-based analytical models of
314VFLUX2, a site-specific thermal diffusivity estimated from vertical diurnal
315temperature signal transport (e.g. Luce et al., 2013), and an estimated
316sediment porosity of 0.5 (fine floodplain sediments).

3172.4 *Water quality monitoring and sampling*

318 Two types of water chemistry data were collected: 1. spatially-
319distributed synoptics covering seepage zones, off-channel ponded areas, and
320main channel locations, and 2. main channel high-resolution time series over
321years 2017-18. Synoptic water samples were collected using 60 mL plastic
322Luer-Lok syringes and filtered through single-use Millex 0.45- μ m Luer-Lok
323filters. There were four synoptic sampling events in total: 1. August 19-22,
3242017; 2. June 21-22, 2018; 3. July 29-August 3, 2018; and 4. September 23-
32525, 2018, with the largest suite of samples collected during event 3, and a
326subset of sample locations visited during other events. Sample water was
327stored in 125 or 250 mL polypropylene bottles, preserved with 2-ml trace
328metal grade HNO₃, and kept in an ice cooler or refrigerated until evaluated
329for dissolved Fe, Mn, As, and Al by either the U.S. Geological Survey Water
330Quality Laboratory or the University of Connecticut Center for Environmental
331Sciences & Engineering Laboratory. Main channel chemical time series of Fe,

332Mn, and Al were collected by grab sample every few days from spring into
333the fall of 2017 and 2018. Time series sample collection locations were
334approximately 1 km downstream of the Coal Creek and East River beaver-
335impacted floodplain zones. Stream water samples were collected daily to
336weekly depending upon snow and ice conditions using an automatic water
337sampler (Model 3700; Teledyne ISCO, NE, USA), with samples pumped via
338peristaltic pump into uncapped 1 L polyethylene bottles. Sample bottles
339were retrieved at regular intervals, with 25 mL aliquots filtered (Pall, NY,
340USA; PTFE; 0.45 μm) and preserved with trace metal grade 12 N HNO_3 until
341analysis. Cation and trace metal concentrations were determined using ion
342coupled plasma mass spectrometry (ICP-MS) (Element 2, Thermo Fisher, MA,
343USA).

344 Field parameters (dissolved oxygen (DO), specific conductivity at 25 °C
345(SpC), and temperature) were typically evaluated at the time of synoptic
346water sample collection with a SmarTroll MP handheld sensor (In-Situ Inc.,
347United States). DO was also tracked over time in summer 2018 in two of the
348larger East River floodplain ponds using MiniDOT loggers (Precision
349Measurement Engineering, Inc., Vista, CA, USA) paired with electrical
350conductivity/pressure loggers (Solinst Levellogger Junior Edge, Solinst Canada
351Ltd, Ontario, CAN). To investigate temporal redox dynamics of beaver return
352flow seepage, a vertical profile of redox potential (Eh) was also collected
353(surface pool, 0.05, 0.1, 0.15, 0.2, 0.25 m depths) at the same focused return

354flow seep mentioned above from June 22 to July 13, 2018 using a custom
355designed logging (1 min increments) redox probe (Paleo Terra, Netherlands).

356 Although thermal infrared is useful for locating surface seepage
357locations, the geometry of the flowpaths that feed those seepage zones, and
358their connection to upgradient water sources, is typically inferred. However,
359near-surface electrical geophysical methods can be used to map flowpaths of
360reduced groundwater, as various redox processes release ions into solution
361increasing the bulk electrical conductivity (EC) of the subsurface (Binley et
362al., 2015). We used a hand carried electromagnetic induction GEM-2
363frequency-domain instrument (Geophex, Ltd.) to evaluate bulk conductivity
364of the near surface. Data were collected in the vicinity the East River return
365flow seeps instrumented with iButton sensors on September 23, 2018 and
366throughout the Coal Creek beaver-inhabited floodplain corridor on
367September 25, 2018. The GEM2 tool was operated over 7 frequencies
368ranging 1,530-93,090 Hz and the expected depth of investigation limit was
369approximately 5 m. Similar to the groundwater/surface water exchange
370study of Ong et al. (2010) we did not invert the data but instead work with
371apparent bulk electrical conductivity (EC) , which was estimated from raw
372(e.g. not smoothed) quadrature data using EMInvertor software (Geophex,
373Ltd.) based on the GEM-2 instrument coil separation (1.66 m).

3743. Results and Discussion

375 A combination of drone-based imaging and ground-based heat tracing,
376 geochemical and geophysical measurements indicated beaver-induced
377 floodplain exchanges create important, and perhaps dominant, transport
378 pathways of natural metals. All data presented below are publicly available
379 from Briggs et al., (2019b, 2019a) and Williams et al., (2019).

3803.1 *Spatial dynamics of geologic seeps and beaver pond return flows*

381 Numerous types of beaver dams, ponds, and return flows were
382 observed along the East River and Coal Creek study reaches, some of which
383 are shown Figure 1b-e. Heat tracing was used to identify spatially
384 preferential channel/floodplain and groundwater connectivity via thermal
385 infrared and FO-DTS technology. Specifically, riverbed interface temperature
386 was recorded with FO-DTS over 6 days in August 2017 along the main East
387 River channel adjacent to the ponded floodplain. Mean temperature along
388 the cables generally ranged from 10.0 to 10.6 °C (full diel range of
389 approximately 8 °C or less), showing subtle warming with downstream
390 distance over the 2.5 km beaver impacted reach (Figure 2). No strong cold
391 anomalies approximating deeper groundwater (approximately 7-9 °C)
392 temperature were observed, indicating discharge of deeper flowpaths to the
393 river is likely not an important process of hydrologic exchange along beaver-
394 impacted section of floodplain. This finding is consistent with the FO-DTS-
395 based hydrogeological characterization of Pai et al. (2017) for a meandering
396 reach immediately downstream of our study reach. Although there are

397several steep, 10's of m high cutbanks into the shale bedrock along the
398reach that might be expected to produce groundwater seepage (Winter et
399al., 1998), shale is typically of low permeability and no substantial
400groundwater discharge was observed from the outcrops over two summer
401field seasons. A few discrete valley wall seepages were located visually/with
402infrared, and groundwater discharge from these was captured by beaver
403ponds before entering the river, as discussed below.

404 Several discrete warm temperature sections are notable in the mean
405FO-DTS record (Figure 2). During retrieval of the FO-DTS cable, we found that
406approximately 6 of these warm anomalies resulted from new beaver dam
407shunts that had been built since the cable was deployed, of the type
408depicted in Figure 1b. These locations are also indicated by large
409temperature standard deviation anomalies, as the cable was exposed in part
410to dynamic air daily temperatures. These high variance zones are spatially
411coupled with slightly cool, less variant temperatures where the cable was
412buried inside the dam materials (Figure 2). The fortuitous real time
413observations of shunt dam building shows channel water diverting structures
414can be built by beaver in just a few days, altering river/floodplain
415connectivity in a substantial and sustained way. Two additional discrete
416sections of FO-DTS cable showed paired warm mean temperatures and low
417standard deviation ('R' in Figure 2). These are interpreted (and field
418confirmed with bed temperature probing) as return flow seepage zones, as
419strong upwelling of even relatively warm water is expected to buffer riverbed

420interface temperature. Both return flow seepage locations were located
421adjacent to major floodplain beaver impoundments. Our FO-DTS results show
422that return flow seepages can be of high enough magnitude to measurably
423alter sediment-water interface temperature in the main channel of larger,
424fast flowing rivers where heat tracing methods are typically challenged to
425locate zones of exchange.

426 Thermal infrared surveys conducted throughout the East River beaver
427reach in summer 2017 and 2018 showed that the floodplain ponds were
428typically warmer than the main channel by afternoon, and that beaver pond
429return flows can be identified as warm anomalies (e.g. Figure 3C), as
430indicated by FO-DTS. In contrast one small beaver pond along the steep
431valley wall of Coal Creek was entirely sourced by a large hillslope spring of
432presumably deeper groundwater (Figure 1d). This spring water was colder
433than the main channel (Figure 3d), demonstrating that not all return flows
434will contribute to warming of channel water in summer, which agrees with
435the finding of Weber et al., (2017) that beavers can enhance thermal
436heterogeneity (cold and warm) in some systems. However, the larger ponded
437areas along Coal Creek floodplain away from the valley wall contained
438relatively warm, diverted channel water, similar to the East River floodplain.
439A more spatially extensive thermal infrared survey conducted along the
440upper East River corridor where the valley is much narrower and steeper,
441and along the bedrock lined, steep Oh-be-Joyful Creek, identified dozens of
442cold groundwater discharges emanating directly from fractured bedrock. A

443subset of these 'geologic' seeps were sampled for chemical comparison to
444the beaver pond return flows, as described in Section 3.2 below.

445 Visual imagery collected by sUAS was integrated to build high
446resolution orthomosaics of the East River (Figure 4) and Coal Creek (Figure
4475) beaver reaches. Surface digital elevation models were also derived from
448the visual imagery and used to infer floodplain surface flow patterns based
449on elevation changes (Supplemental Figures A1, A2). Although not
450attempted for this study, such structure-from-motion drone imaging products
451are likely to be useful for 'fill and spill' numerical flow modeling of ponded
452areas. The 2017 East River orthomosaic demonstrates the extensive
453saturated floodplain area induced by beaver-induced shunting of channel
454water (Figure 4a, b). It appeared that just 2 shunts placed at strategic
455locations, namely at the confluence of a river oxbow and along a river side
456channel, were responsible for the majority of diverted channel water over
457both summer seasons (Figure 4 a). These shunts were less effective in
458July/August 2018 due to a lower flow condition causing widespread draining
459of the floodplain ponded areas (Figure 4c), though the floodplain morphology
460appeared comparable to 2017. A rain event the day before the 2017 sUAS
461mapping mobilized fine sediments and the resulting turbidity was used as a
462natural qualitative tracer of advective flow connectivity through the linked
463ponded systems (Figure 4b). These flow patterns indicate preferential
464pathways through more stagnant ponded areas.

465 The pond systems generally terminated near a large meander bend of
466the river, where clusters of beaver pond return seeps transferred water back
467to the channel (Figure 4b, c). No prominent surface return flows were noted
468in summer 2017 or 2018 along the East River floodplain section. However, a
469survey in later September 2018 showed that at the lowest channel flow
470condition beaver were able to build several spanning dams across the East
471River, diverting more water into the floodplain, refilling and overflowing the
472ponded areas and creating numerous overland return flows. It may be that
473alluvial mountain rivers of similar large size to the East River go through a
474natural beaver diversion cycle: 1. Large spring snow melt pulses damage or
475destroy the previous year's dam structures (also shown by Briggs et al.,
476(2013)); 2. In early summer river discharge recedes but river stage is still
477relatively high and shunts are effective to divert water to the floodplain, but
478channel-spanning dams cannot yet be constructed (Figure 1b); 3. In mid-
479summer, channel flow drops farther (typically by a factor of 10x from spring
480peak at the East River, i.e. USGS gage 09112500) and the shunts are less-
481effective but channel spanning dams cannot yet be built, causing a recession
482of floodplain pond levels (Figure 4c); 4. At the lowest flows in early fall,
483spanning dams are built, refilling the ponded areas before winter. Longer
484term, higher-frequency monitoring is needed to explore these temporal
485dynamics. In contrast to the East River, channel spanning dams were
486observed along the Coal Creek system across 2017 and 2018 summer and
487early fall seasons (Figure 5a). Remarkably, each of four successive dams

488temporarily diverted almost the entirety of channel flow into the adjacent
489floodplain in 2018 (Figure 5a, c, d), and this diverted water returned to the
490downstream channel in a series of surface and subsurface return flows.
491Discharge from two of the larger surface return flows was measured at 173
492and 346 m³/d (with mobile weirs) in August 2018, representing a large input
493of sub-oxic water back to the main channel. Extensive Fe-oxide staining was
494visible along the floodplain ponded areas (rust colors, Figure 5c), but oxide
495deposition was not associated with the smaller groundwater spring-fed
496beaver pond. A downstream ponded floodplain area captured another
497groundwater discharge originating from a road culvert on the hillslope above
498the floodplain (Figure 5d), and this oxic groundwater mixed with reduced
499floodplain water before entering the channel in a series of subsurface
500seepages. Beaver dam capture of discrete hillslope groundwater discharge
501was also noted at multiple locations along the East River, indicating that
502floodplain ponds should be considered in groundwater/surface water
503exchange studies that are typically focused on hyporheic exchange alone.
504Google Earth imagery from 1999, 2005, and 2012 of the Coal Creek reach
505showed similar (to 2018) floodplain pond morphology and the existence of
506channel spanning dams diverting large portions of streamflow (Supplemental
507Figure A3), indicating 'disturbance' caused by beaver inhabitation may
508create a relatively stable new floodplain exchange dynamic.

5093.2 *Dissolved chemistry of beaver-induced floodplain exchanges and*
510 *geologic seeps*

511 Earlier work has indicated the potential for beaver impoundments to
512 expand zones of reducing conditions in saturated soils (Cirimo and Driscoll,
513 1993; Naiman et al., 1988). Recently, NRZs have been identified in other
514 Colorado floodplain systems as key locations of nutrient transformation
515 (Boye et al., 2017; Dwivedi et al., 2018) and contaminant accumulation
516 (Janot et al., 2016). However, although NRZs have strong, spatially
517 compressed redox gradients, they are not all likely to function as hotspots of
518 reaction influential to the larger floodplain system chemistry, or ecosystem
519 'control points.' Reducing conditions can develop locally due to enhanced
520 organic carbon availability and/or residence time (Boano et al., 2010), but
521 spatially-compressed redox gradient alone does not indicate mass flux of
522 reduced chemical species. To influence mixed river water metals
523 concentrations, NRZs must also have appreciable advective exchange with
524 the channel. Our sUAS-based surface mapping and return flow observations
525 indicate beaver-induced flowpaths may dominate river-floodplain advective
526 flux compared to other types of lateral exchange in these systems (Figure
527 a), but a more quantitative picture is developed with chemical analysis.

528 Synoptic chemical samples were collected at a combination of main
529 channel (29 samples), beaver pond (14 samples), beaver pond return flow
530 (17 samples), and geologic seep (14 samples) locations; although all
531 parameters (DO, SpC, Mn, Fe, Al, As) were not always evaluated for each

532sample. Although floodplain ponds were relatively easy to physically access,
533sample filters clogged quickly there, practically limiting pond sample
534numbers. Spot measurements at the time of water sample collection showed
535return flows had the lowest median DO concentration at 47% saturation
536(Figure 6a). However, overall return flows ranged from fully anoxic to fully
537saturated in DO. This large range can be explained by return flows being
538comprised of both surface and subsurface flowpaths, with the latter
539generally of considerably lower DO saturation. Surprisingly, most pond
540samples were super-saturated in DO, owing to abundant observed primary
541production (filamentous algae) in the shallow open pools, as all samples
542were taken during daytime hours. The temporal records from the two major
543East River floodplain beaver ponds show a more complete story, with large
544swings from daytime DO super saturation (e.g. >11 mg/L DO diurnal swings)
545to nearly anoxic conditions overnight, suggesting a system with strong
546continual aerobic respiration (Figure 7b). Elevation-corrected DO saturation
547was estimated with the Benson and Krause Equations (US Geological Survey,
5482011). Oxygen is the master variable that controls redox condition
549(Zarnetske et al., 2012), so strong daytime photosynthesis signal of beaver
550ponds can impart a highly dynamic redox signal onto return flowpaths that
551are otherwise suboxic. The DO time series data indicate that our daytime
552pond grab samples for dissolved metals may underestimate daily average
553levels, as night time suboxic conditions would be expected to enhance metal
554concentrations. As the 2018 summer progressed, the pond became suboxic

555 though extensive pond algae was still observed (Figure 7b), so it is possible
556 net aerobic respiration increased during this period, and/or advective
557 circulation of the ponds decreased at lower water levels.

558 The SpC of beaver pond return flows showed the largest median
559 conductivity at 351.0 $\mu\text{S}/\text{cm}$, with the high end of that range driven by
560 seepages (Figure 6b). This result indicated the potential for subsurface
561 return flow pathways to be mapped with electromagnetic imaging due to
562 enhanced bulk EC, as described below. Ponds showed the largest total range
563 as SpC driven in part by pre-sampling precipitation events and mixing with
564 valley wall groundwater discharges. Other types of measured groundwater
565 discharge, predominantly from fractured bedrock, also showed a large range
566 in SpC but the lowest median value at 158.8 $\mu\text{S}/\text{cm}$.

567 Beaver-induced flows were most distinct from other river corridor
568 water sample types in respect to dissolved concentrations of Fe and Mn
569 (Figure 6 c,d; Figure 7c). Return flows averaged (median) 1120.0 and 210.6
570 $\mu\text{g}/\text{L}$ for Fe and Mn, respectively, with the maximum Fe value of 14,260.0
571 $\mu\text{g}/\text{L}$ collected in August 2017 at the major East River return flow seepage
572 instrumented with a redox profiler. For contrast, the median Fe and Mn
573 concentrations ($\mu\text{g}/\text{L}$) in the other three types of samples are: channel (169.1
574 Fe / 4.7 Mn), beaver ponds (366.8 Fe / 19.2 Mn) and geologically controlled
575 groundwater (54.9 Fe / 1.3 Mn). As floodplain beaver pond water is
576 dominated by channel diversions, with some discrete hillslope groundwater

577inflow, metal concentrations are clearly increased by beaver-induced
578hydrologic exchanges.

579 While it is not uncommon to find high concentrations of natural metals
580in reduced floodplain soil porewater (Schulz-Zunkel and Krueger, 2009), what
581makes beaver pond return flows unique is that they also show strong
582advective flux. Hyporheic exchanges in larger river systems often may not
583substantially impact mixed river solute transport, particularly at the reach-
584scale (Wondzell, 2011). However, in the East River system dissolved Mn
585concentrations collected in 4 surveys over a year always increased in mixed
586main channel water along the beaver-impacted floodplain (Figure 4a).
587Background concentrations of Mn were substantially higher in the mine-
588impacted Coal Creek reach, and although channel sampling was more
589limited, large increases in concentration were observed over just a few
590hundred meters in the zone of return flows (Figure 5a). Plotting Fe vs Mn for
591all samples clearly demonstrates how return flows from beaver ponds
592dominate the anomalously high concentrations observed for both species,
593and although the ratio of the metals differed, Fe concentrations were almost
594always dominant (Figure 7c) consistent with Fe being preferentially elevated
595in comparison to Mn in the vast bulk of geologic materials. The mobility of As
596and Al was enhanced by beaver-induced floodplain exchanges (Figure 6e, f),
597as discussed in Section 3.3.

598 The spot DO measurements at East River beaver pond return flow
599seepages all showed varied degrees of suboxic condition (Figure 6a, Briggs
600et al., 2019a), though temporal redox fluctuations are not clear in these
601sparse sampling events. However, the redox potential profile collected
602directly within the return flow seep (shown flowing in Figure 3d) had
603systematic Eh shifts at all depths at daily to weekly timescales (Figure 7d).
604Overall there was a transition toward strong reducing conditions from June
60522 to July 11, 2018, except in the surface seepage pool where the probe was
606likely exposed to air periodically. The reducing shift likely results from the
607observed decreased seepage rates over time. Total flow from the seepage
608was physically measured to be 1464 L/d in late June but was too low to be
609reliably captured with the surface weir in late July, a reduction explained by
610the observed recession of the upgradient pond level during this period
611(decreased lateral hydraulic gradient). Vertical seepage rates measured over
61270 d in 2017 using iButton temperature sensors installed in this seep show
613coordinated short (daily) and longer-term flux patterns, also indicating that
614seepage redox chemistry (and associated metal concentrations) is likely to
615fluctuate over time (Figure 7a). As discussed in detail by Briggs et al. (2013),
616reactive mass flux beaver pond return seepages are not likely to occur at
617highest metal concentration but when fluid flux and concentration (typically
618inversely related) are optimally balanced. Therefore, higher flux surface
619return flows of lower metal concentration may be more important to river
620chemical dynamics than strongly reduced focused seepages. For example,

621the predominant East River return flow seepage transferred approximately
62210 g/d of dissolved Fe²⁺ to the main channel at times during this study,
623while the larger Coal Creek surface return flow transferred approximately
624218 g/d Fe²⁺.

625 In general, the focused return seepage water was less reduced toward
626the land surface along the redox profiler, indicating some vertical diffusive
627exchange with surface oxygen, and/or a convergence with oxic subsurface
628flowpaths at the seepage zone. During redox probe installation it was clear
629that beneath approximately 10 cm of fine sediments the focused seepage
630zone sediments were composed of higher permeability sands and gravels. As
631has been observed for numerous other river corridor seepage types, the
632distribution of spatially focused return flow seepages is likely controlled by
633existing heterogeneous floodplain geologic deposits where relatively coarse
634alluvium creates conduits of hydrologic exchange.

635 The areal 'footprint' of sub-oxic return flowpaths was mapped from the
636land surface using electromagnetic imaging. Higher frequencies of the GEM2
637tool should represent more shallow subsurface bulk EC dynamics, so raw
638data from the highest four frequencies (of 7 total frequencies) were
639arithmetically averaged for this analysis, as the lowest 3 frequencies were
640found to have reduced sensitivity in these systems. The resulting map of
641electrically-conductive subsurface anomalies below a larger beaver pond at
642the East River indicated a swath of reduced water 10's of meters across

643 flowing in the shallow subsurface toward the river, some of which discharges
644 at the focused seep where the redox probe was installed (Figure 4c). These
645 reactive flowpaths also source the diffuse seepage zones along the main
646 channel margin, including the Fe-rich side channel shown in Figures 1d) and
647 e). In general, electromagnetic imaging data collected throughout the
648 channel area and over the opposite bank floodplain where there was no
649 beaver activity did not indicate extensive subsurface plumes of metal-
650 impacted water (Figure 4c). Vertical seepage rates along the channel margin
651 were slow, typically less than 0.2 m d^{-1} over the 2017 period monitored with
652 vertical iButtons (Figure 7a), but spatially extensive enough to drive the DO
653 content of the side channel surface water down to an average of 54%
654 saturation in mid-day during the summer 2018. Vertical seepage rates during
655 the same period in 2017 for the focused beaver return flow seepage zone
656 were stronger, ranging up to 0.6 m/d . Diffuse seepage rates at all 4 locations
657 showed coordinated short-term shifts to downward flow, likely due to higher
658 event flows in the channel. In contrast, discharge from the focused bank
659 seep showed a different temporal pattern that is likely driven by beaver pond
660 stage dynamics, and not directly impacted by channel flow.

661 A larger area was imaged along the Coal Creek, revealing 3 major 'hot
662 spots' of increased shallow bulk EC (Figure 4b). The two upstream zones are
663 adjacent to the main floodplain ponded areas in the vicinity of observed
664 return flow seepages. This result agrees with the East River imaging, in that
665 although seepages may be highly focused in space at the land surface, they

666are fed and underlain by larger subsurface plumes of reduced metals. Of
667note, the downstream surface return flows, although enriched in Fe and Mn
668compared to channel water, do not create any extensive electromagnetic
669anomaly. This may be surprising given the extensive visible Fe staining along
670the creek sediments in this area (Figure 5c), but the GEM2 tool is sensitive to
671the upper several meters of earth material, and therefore this result
672indicates that the surficial return flows are not underlain by extensive
673subsurface reduced plumes. Much of the Fe oxides visible in this area may
674be precipitating from the nearby upstream highly reduced return flow
675subsurface seepages, and/or from the moderate concentrations of reduced
676Fe measured in the surface return flows. Further downstream, a channel
677spanning dam diverts channel water to both the right and left bank
678floodplain areas (Figure 5d). Flowpaths along the right bank appeared to stay
679on the land surface, remained oxic, and there was little enhancement of
680subsurface bulk EC. However, the downstream left floodplain area was highly
681reduced, mixing a valley wall groundwater seep with diverted channel water
682that returned to the stream in a series of seepages. Fe staining was
683prevalent in this area, and the floodplain pools and shallow subsurface highly
684electrically conductive (Figure 5b).

685 The main channel chemical time series (Fe, Mn, Al) were collected
686approximately 1 km downstream of each beaver-impacted reach (Figure 8).
687In total 299 samples were collected at the East River and 340 samples
688collected at Coal Creek over the 2017/2018 period. Main channel Fe

689 concentrations were typically measurable, indicating persistent inflow from
690 reduced seepages. There was a bimodal pattern with early and late season
691 peaks in concentration (over 50 ppb) at the East River that may be tied to
692 the strong early and late season beaver-induced floodplain connection
693 mentioned above (Figure 8a). Mn was generally quite low or below the
694 detection limit, except for a few spikes coinciding with Fe highs. When East
695 River Fe and Al are plotted against each other there is no strong proportional
696 relation (Figure 8b), indicating other processes in addition to reduced return
697 flows drive Al concentration dynamics. This may be explained in part by the
698 high concentrations of Al observed at the floodplain valley wall geologic
699 groundwater seepages, such that some combination of groundwater
700 discharge and beaver-induced floodplain exchange influences downstream Al
701 concentration. A bimodal temporal pattern of main channel Fe concentration
702 was also observed at Coal Creek (Figure 8c), where Mn concentrations were
703 generally much higher than at East River and better coupled with Fe. A
704 strong proportional relation was observed between Fe and Al at Coal Creek
705 indicating that for that system reduced return flow seepages may drive Al
706 mobility (Figure 8d). Unlike the East River, Al concentration in the local Coal
707 Creek beaver-impacted reach hillslope groundwater was low, based on
708 limited data.

709 3.3 Metal oxide deposition at beaver pond return flows

710 Oxides and hydroxides (referred to here by the general term “oxides”)
711 of Mn and Fe metals are often associated with groundwater seepage zones

712(Boano et al., 2014; Gandy et al., 2007) and characteristic red staining of
713surface sediments is frequently used to visually locate points of sub-oxic
714seepage (e.g. Figure 1e; Figure 5c,d). Similar to engineered geochemical
715barriers using zero-valent Fe (McCobb et al., 2018), metal oxides function as
716a sorption sink for a host of dissolved contaminants toxic to humans and
717aquatic life. In watersheds, such as Coal Creek that are impacted by mine
718water drainage, Mn oxides have been shown to sorb and co-precipitate
719cobalt, nickel, and zinc in high concentrations (Jenne, 1968). Fe oxides have
720also been shown to substantially reduce these contaminants as groundwater
721flows through the streambed, and to be a strong sink for arsenic (Nagorski
722and Moore, 1999). In zones of uranium contamination, adsorption to biogenic
723Fe oxides can strip hexavalent uranium (U(VI)) from groundwater
724(Katsoyiannis, 2007) before discharge to surface water. Fe oxides have also
725been shown as important sorption sites for perfluorooctane sulfonate (PFOS)
726(Johnson et al., 2007), a contaminant of major emerging concern (Banzhaf et
727al., 2016). However, dynamic dissolution of oxides under dynamic reducing
728conditions of beaver-impacted floodplain soils can mobilize previously
729sequestered contaminants along with the dissolved metals.

730 Along the river corridor, solid grain Fe and Mn is typically found in
731glacial sediment grains, alluvial sediments, and mine tailings. Reduction to
732soluble form under suboxic condition mobilizes the metals to travel with
733hyporheic flow, groundwater, or as this study has shown, in surface return
734flows and subsurface seepages. Widespread Fe staining below return flow

735 discharge points along the East River and Coal Creek corridors visibly
736 indicates how beaver activity can greatly alter metal oxide dynamics in
737 alluvial systems (Figure 5c, d). For example, a side channel along the East
738 River adjacent to the return seepages that was instrumented with iButtons
739 and the redox probe, was shown to collect reduced water loaded with natural
740 metals (Figure 1e), and precipitate oxides as this water exchanged gas with
741 air and advectively mixed with the main channel.

742 Arsenic and aluminum concentrations were predominantly higher in
743 the mine-impacted Coal Creek return flow samples as compared to the East
744 River, averaging 6.3 and 10.1 $\mu\text{g/L}$, respectively. These concentrations are
745 approximately 2x higher than that observed in the diverted channel water,
746 suggesting the mobility of these contaminants is tied to the dissolution of
747 floodplain metal oxides. Considering the importance of metal oxides to a
748 host of abiotic and biotic processes, beaver pond return flows of reduced
749 water could be recognized as ecosystem control points (Bernhardt et al.,
750 2017), and deserve similar research attention to more commonly studied
751 mechanisms of river to floodplain hydrologic exchange. Several western USA
752 alluvial river corridors with similar morphology to East River have
753 contemporary U(VI) contamination concerns resulting from legacy floodplain
754 mine tailings (Curtis et al., 2006; Naftz et al., 2018), and it has been shown
755 that mobility of U(VI) is directly tied to Fe oxide dynamics in NRZs (Bone et
756 al., 2017; Davis et al., 2006). In such systems the return of beaver, or human

757simulation of their dam construction using dam analogues, may result in
758undesirable transport of contaminants.

7594. **Conclusions**

760 Enhanced river/floodplain hydrologic connection has been shown to
761increase river corridor evapotranspiration and net carbon uptake (Missik et
762al., 2018), but the impact of beaver-induced floodplain water flux on the
763mobility of river corridor metals has been largely under-characterized. In the
764two alluvial systems studied here, we observed high-flow active shunting of
765stream water onto adjacent floodplains, greatly expanding the volume of
766saturated floodplain sediments with strong hydrologic connectivity to the
767channel. Land surface and subsurface beaver pond return flows contained
768high concentrations of dissolved Mn and Fe, redox-sensitive metals that are
769highly influential to a multitude of biogeochemical and abiotic processes.
770Dissolution of solid phase floodplain sediment Mn and Fe oxides can provide
771an advective pathway for contaminant transport, particularly in mine-
772impacted watersheds. In contrast to episodic overbank river flow events or
773slower-exchanging meander bend flowpaths, beaver-induced exchanges can
774provide strong, persistent river-floodplain connectivity and conduits for metal
775mobility. As beaver return to alluvial floodplain systems across north
776America, active human management will likely need to consider system-
777specific consequences of enhanced exchange with suboxic floodplain waters,

778to be balanced against numerous desirable hydroecological and restorative
779outcomes.

780

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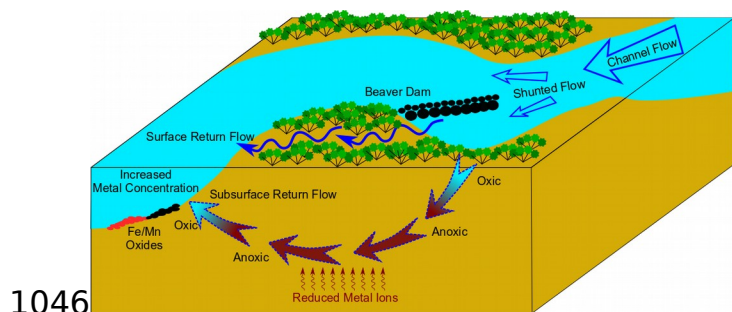
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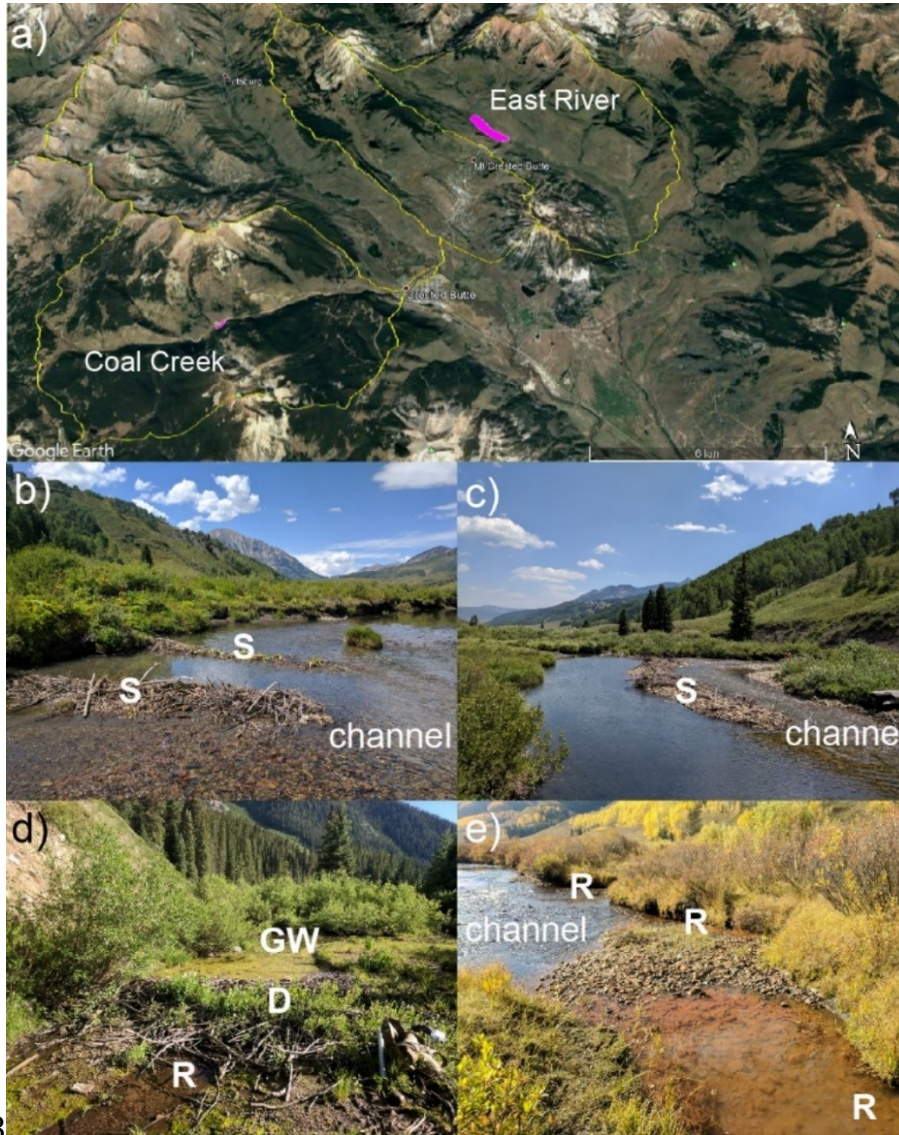
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1044

1045 **Graphical Abstract**



1047 Figures

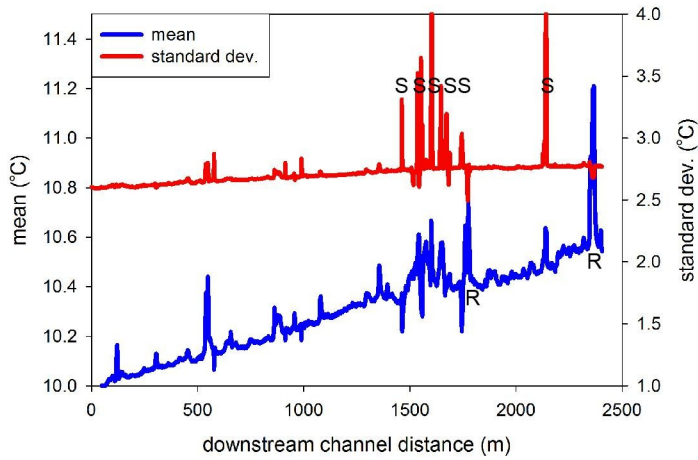


1048

1049 Figure 1. Panel a) shows the sub-watersheds of the East River SFA, where
1050 the Coal Creek and East River beaver-impacted study floodplain sections are
1051 highlighted in pink. In spring and summer, beavers construct a series of
1052 dams at the East River to 'shunt' (S) large volumes of channel water onto the
1053 adjacent floodplain (panels b,c). Discrete hillslope groundwater (GW) springs
1054 may be directly captured by small beaver dams (D) (panel d) before draining

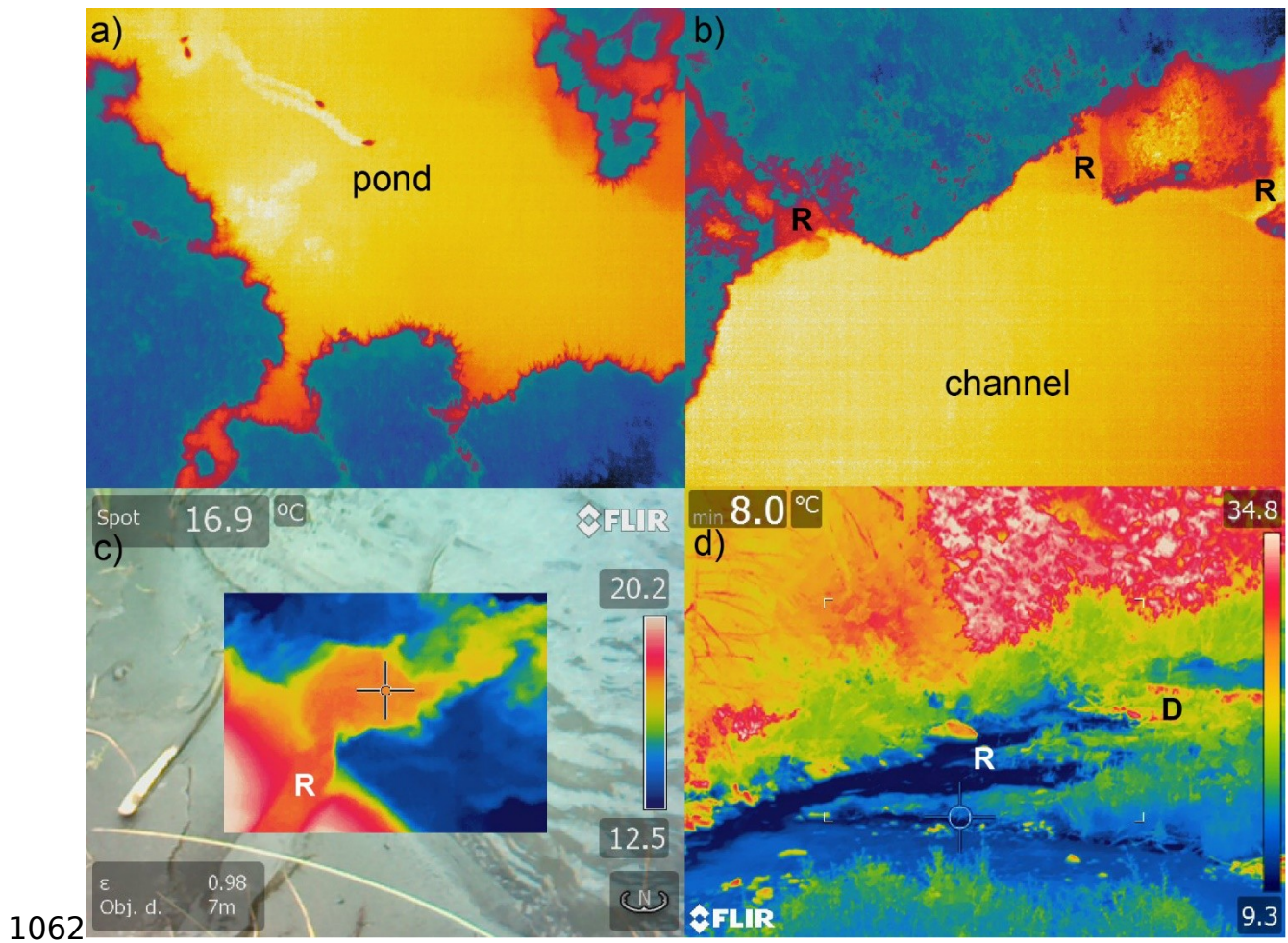
1055to the main channel, while beaver pond return flow seeps (R) are typically
1056warmer and lower in oxygen in summer (panel e).

1057



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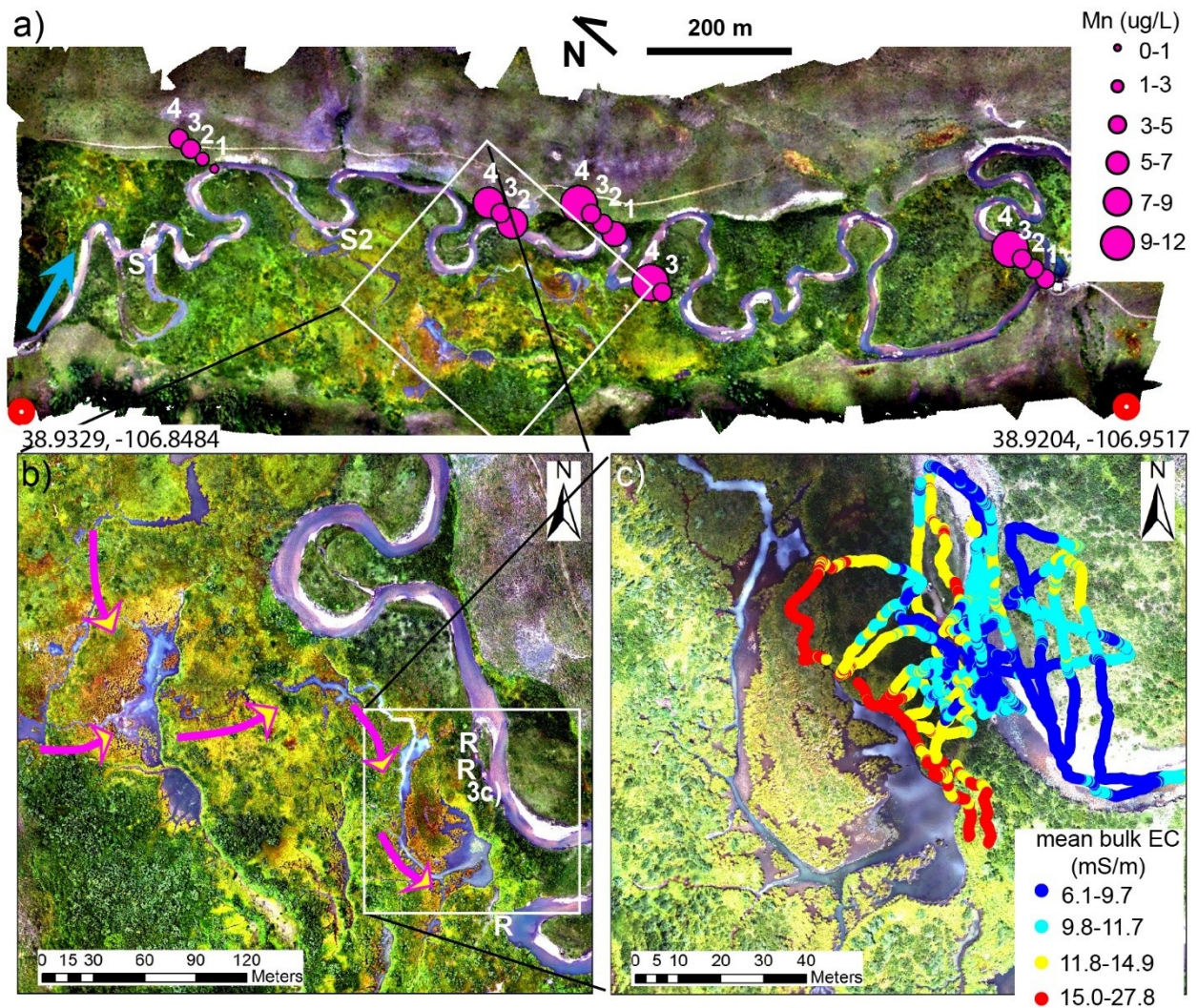
1059Figure 2. The 6-day mean and standard deviation of temperature along the
1060East River fiber-optic cable showing influence of 'shunt' beaver dam
1061construction (S) and shallow, warm beaver pond return flow seepage (R).



1062
 1063 Figure 3. Thermal infrared imaging collected by drone of floodplain beaver
 1064 ponds show relatively warm ponded areas, indicating by the hot colors in
 1065 panel a), and beaver pond return flow seepage is shown by relative color
 1066 scale in panel b), and close up through handheld imaging in panel c).

1067 Hillslope springs captured by floodplain beaver ponds collect relatively cold,
 1068 deeper groundwater that discharges to the main channel after mixing with
 1069 floodplain water (panel (d)).

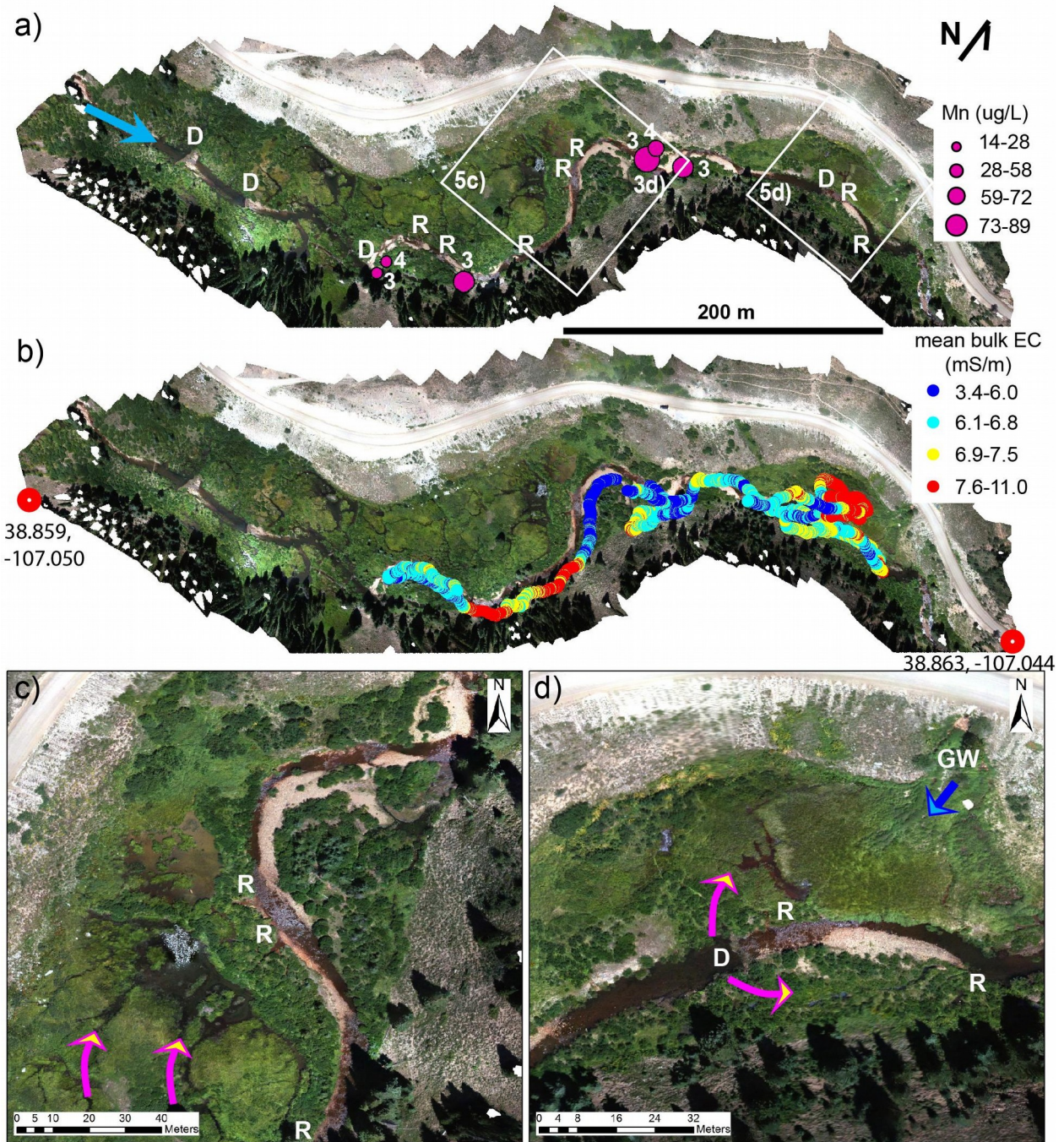
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1071

1072 Figure 4. Panel a) displays the full size orthomosaic generated from 2017
 1073 drone imagery collected along the East River beaver impacted reach (north
 1074 direction rotated left). River flow is left to right, and the two main beaver
 1075 shunts are marked (S1, S2) that push channel water onto the adjacent
 1076 floodplain. Main river channel dissolved manganese concentrations are
 1077 shown for samples collected on: 1. August 2017, 2. June 21, 2018, 3. July 30,
 1078 2018, and 4. September 23, 2018. Panel b) shows an enlarged image of the
 1079 2017 imagery of the more prominent floodplain beaver ponds and major

1080return flow seeps (R), including the approximate location of the infrared
1081image of Figure 3c). General surface flow patterns are shown with yellow
1082arrowheads as inferred from fine, light colored sediment transport following
1083a rain event. Panel c) is an enlarged image of the 2018 drone-based
1084orthomosaic showing lower pond water levels. Electromagnetic imaging
1085transects indicate shallow subsurface plumes of reduced water (higher bulk
1086conductivity) extending from the ponded area toward the main channel.

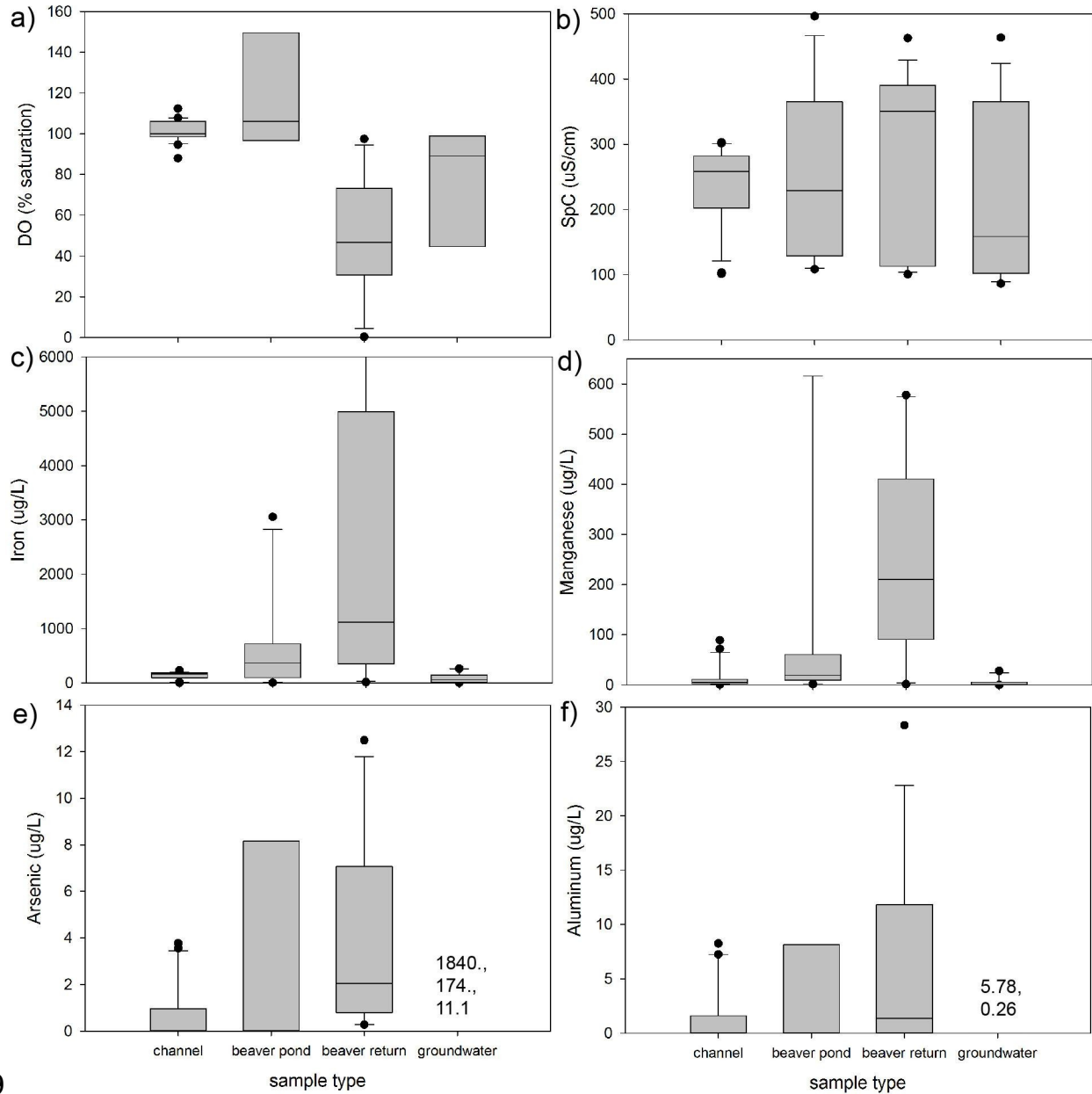


1087

1088 Figure 5. Panel a) shows the full orthomosaic generated from 2018 drone
 1089 imaging collected along the Coal Creek beaver impacted reach. Stream flow
 1090 is left to right and channel spanning beaver dams (D), return flows (R), and
 1091 the locations of Figure panels 3d), 5c), and 5d) are marked. Main channel

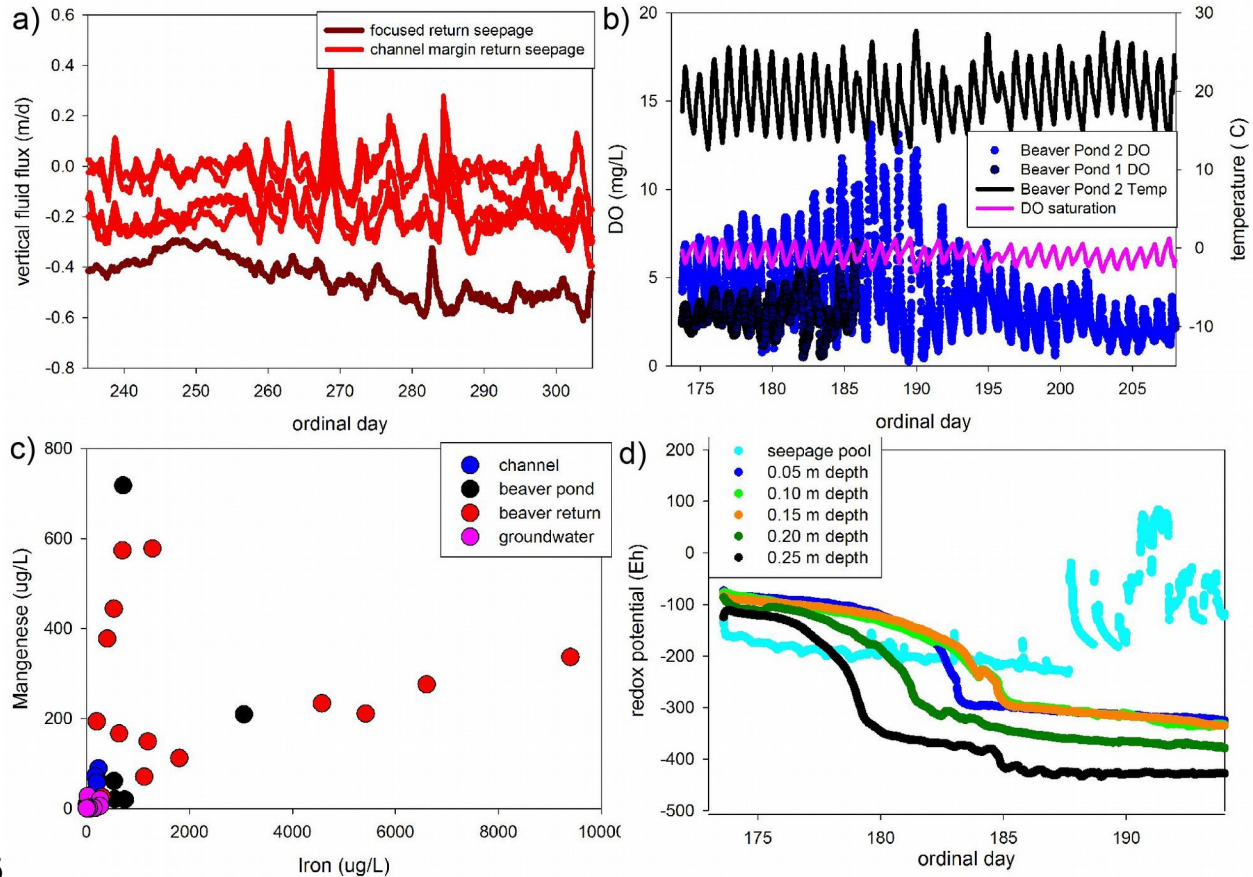
1092 dissolved manganese concentrations are shown for two sampling events on:
1093 3. August 2, 2018, and 4. September 25, 2018. Panel b) covers the same
1094 spatial extent as a), and shows electromagnetic imaging transects data.
1095 Panel c) displays an enlarged view of the larger floodplain ponds, and Panel
1096 d) shows a zoomed view of the most downstream channel spanning dam and
1097 resulting floodplain diversions.

1098



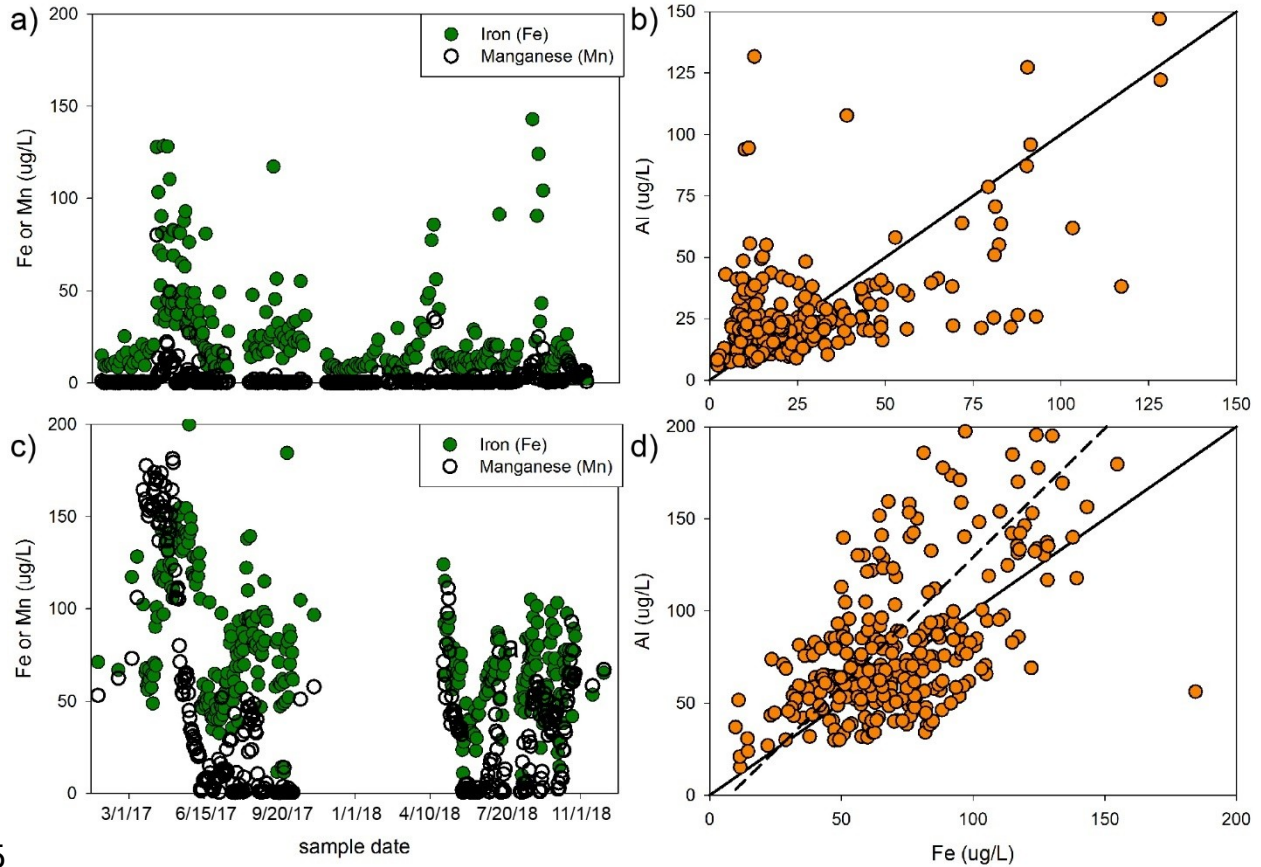
1099

1100 Figure 6. Water samples of source type: a) dissolved oxygen, b) specific
 1101 conductivity, c) iron, d) manganese, e) arsenic, and f) aluminum. The vertical
 1102 box indicates the interquartile range and the dots are outliers. For plots e)
 1103 and f) all groundwater samples were below the respective detection limits
 1104 except for the discrete values listed. The full chemical dataset is listed by
 1105 sample in the Supplemental Material.



1106

1107 Figure 7. Panel a) displays summer/fall 2017 vertical return flow seepage
 1108 rates, while panel b) shows measured dissolved oxygen (DO), temperature,
 1109 and theoretical oxygen saturation for East River beaver ponds. Panel c)
 1110 shows a plot of dissolved iron vs. manganese for all water samples of varied
 1111 type; a sample of 14260.0 $\mu\text{g/L}$ Fe and 472.5 $\mu\text{g/L}$ Mn collected at the same
 1112 location of the redox profile is not displayed. Panel d) displays multi-depth
 1113 redox potential (Eh) monitored directly at the discharge point over time at a
 1114 major East River return flow seep (shown in Figure 3b).



1115

1116 Figure 8. Time series of iron, manganese, and aluminum collected over 2017-
 1117 2018 1 km downstream of the distal end of the East River (panels a,b) and
 1118 Coal Creek (panels b,c) beaver-impacted reaches. The solid line in panels b)
 1119 and d) indicates a 1:1 relation while the dashed line in panel d) indicates the
 1120 best linear fit to the data ($R^2=0.52$), no significant linear relation was found
 1121 for the data in panel b).