



**Effects of Plot Size, Stand Density and Scan Density on the Relationship between Airborne Laser Scanning Metrics and the Gini Coefficient of Tree Size Inequality**

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1 **Title:**

2 Effects of Plot Size, Stand Density and Scan Density on the Relationship between Airborne  
3 Laser Scanning Metrics and the Gini Coefficient of Tree Size Inequality

4 **Authors:**

5 Syed Adnan \*(1), Matti Maltamo (1), David Coomes (2), Rubén Valbuena (2)

6

7 **Affiliations:**

8 (1) University of Eastern Finland. Faculty of Forest Sciences. PO Box 111 Joensuu,  
9 Finland; [adnan@uef.fi](mailto:adnan@uef.fi); [matti.maltamo@uef.fi](mailto:matti.maltamo@uef.fi).

10 (2) University of Cambridge, Department of Plant Sciences. Forest Ecology and  
11 Conservation. Downing Street, CB2 3EA Cambridge, UK. [dac18@cam.ac.uk](mailto:dac18@cam.ac.uk);  
12 [rv314@cam.ac.uk](mailto:rv314@cam.ac.uk).

13 \*Corresponding author.

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**28 Abstract**

29 The estimation of Gini Coefficient (*GC*) of tree sizes using airborne laser scanning (ALS) can  
30 provide maps of forest structure across the landscape, which can support sustainable forest  
31 management. A challenge arises in determining the optimal spatial resolution that maximizes  
32 the stability and precision of *GC* estimates, which in turn depends upon stand density or ALS  
33 scan density. By subsampling different plot sizes within large field plots, we evaluated the  
34 optimal spatial resolution by observing changes in *GC* estimation and in its correlation with  
35 ALS metrics. We found that plot size had greater effects than either stand density or ALS  
36 scan density in the relationship between *GC* and ALS metrics. Uncertainty in *GC* estimates  
37 fell as plot size increased. Correlation with ALS metrics showed convex curves with maxima  
38 at 250-450 m<sup>2</sup>, which thus was considered the optimal plot size / spatial resolution. By  
39 thinning the density of ALS point cloud, we deduced that at least 3 points·m<sup>-2</sup> are needed for  
40 reliable *GC* estimates. Many nationwide ALS scan densities are sparser than this, which may  
41 be unreliable for *GC* estimation. Ours is a simple approach for evaluating the optimal spatial  
42 resolution in remote sensing estimation of any forest attribute.

**43 Key words**

44 structural heterogeneity; spatial resolution optimization; sample size optimization; forest  
45 structure; LiDAR

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## 50 1. Introduction

### 51 1.1 *The Gini Coefficient as an Indicator of Forest Structural Heterogeneity*

52 Forest structural characteristics are widely used in the development of sustainable  
53 management plans designed to protect habitats while carrying out forestry operations (Upton  
54 and Fingleton, 1985; Pommerening, 2002; Motz et al., 2010; Vihervaara et al., 2015;  
55 Valbuena et al., 2016). Management can be designed to emulate natural dynamics (Oliver and  
56 Larson, 1990; Buongiorno et al., 1994; Lähde et al., 1999; Pukkala et al., 2016), by studying  
57 how silvicultural operations affects forest structure locally (Humphrey et al., 2000; Valbuena  
58 et al., 2013a; Robles et al., 2016).

59 Forest structure is often characterized by stem diameter distributions (O'Hara and Gersonde,  
60 2004; McElhinny et al., 2005). If a single concise indicator of size inequality is desired, there  
61 are many available, including Shannon or Simpson indices (Neumann and Starlinger, 2001;  
62 Sterba and Ledermann, 2006; O'Hara et al., 2007; Lei et al., 2009) or variance-based metrics  
63 (Staudhammer and LeMay, 2001). Recent research has highlighted the effectiveness of the  
64 Gini coefficient (i.e. *GC*, Gini, 1921) for assessing the structural diversity (Lexerød and Eid,  
65 2006a; O'hara et al. 2007; Duduman, 2009; Valbuena et al., 2012, 2013a). Originally  
66 developed for evaluating inequality in income distributions (e.g., Hvistendahl, 2014), *GC* has  
67 been applied to a variety of fields, such as healthcare (Asada, 2005) or land use (Zheng et al.,  
68 2013). In plant sciences, it has been employed to evaluate size inequality (Weiner and  
69 Solbrig, 1984). It has also been applied to forest ecosystems (Weiner and Thomas, 1986), to  
70 quantify structural diversity (Knox and Peet, 1989), analyse competition (Lundqvist, 1994;  
71 Cordonnier and Kunstler, 2015), or successional stages (Valbuena et al., 2013a). Comparative  
72 studies indicate that *GC* is the best index for characterizing diameter distributions, providing a  
73 logical ranking of different stand types (Lexerød and Eid, 2006a; Valbuena et al., 2012), so  
74 that forest may be stratified according to their structure (Bollandsås and Næsset, 2007). It can

75 also be used to observe the effects of different management regimes (Bourdier et al., 2016;  
76 Pukkala et al. 2016; Valbuena et al., 2016). For these reasons, estimation of  $GC$  is the focus  
77 of this article.

78 When used in forest science,  $GC$  evaluates size inequality of trees growing in a vicinity  
79 (Weiner, 1990). For a patch of forest containing  $n$  trees, within which the  $i^{\text{th}}$  and  $j^{\text{th}}$  tree have  
80 basal areas of  $g_i$  and  $g_j$  respectively,  $GC$  is calculated as (Glasser, 1962):

$$81 \quad GC = \frac{n}{(n-1)} \frac{\sum_{i=1}^n \sum_{j=1}^n |g_i - g_j|}{2n^2 \bar{g}} \quad (1)$$

82 Therefore,  $GC$  is a statistical measure of relative dispersion, which is equivalent to half of the  
83 relative mean absolute difference (Valbuena et al, 2017: appendix A3), and it ranges between  
84 0-1, zero representing perfect equality and one being maximum inequality (Gini, 1921).  
85 Hence  $GC$  describes the shape of tree-size distributions (Valbuena et al., 2016) and is  
86 influenced by competitive interactions among trees (Cordonnier and Kunstler, 2015).  
87 Valbuena et al. (2012) demonstrated that the  $GC = 0.5$  can be considered as a boundary  
88 between even-aged and uneven-aged stand structures.  $GC$  values far below 0.5 indicate a  
89 unimodal “normally distributed” size structure, which are commonly found in even-aged  
90 stands that are self-thinning (e.g. Coomes and Allen, 2007). Values close to 0.5 indicate  
91 irregular size distributions (Duduman, 2009), while values much greater than 0.5 represent  
92 “reverse-J” stand structures (Valbuena et al., 2013a).

### 93 *1.2 Influence of Plot Size in Measurements of Forest Structure*

94 Sample plots used for measuring plant communities are usually rectangular or circular in  
95 shape (Whittaker, 1972; Kent and Coker, 1992), with a wide range of possible plot sizes from  
96 fine to coarse scales (Chytrý and Otýpková, 2003). As the effects of plot size decrease with  
97 increasing size of a plot (David and Mishriky, 1968; Barbeito et al., 2009), an optimal size

98 must be chosen compromising the acquisition of a field plot large enough to obtain a stable  
99 measure of forest structure, but no larger than necessary because of the costs involved  
100 (Otypková and Chytrý, 2006). Structural diversity depends on the spatial resolution at which  
101 an index is evaluated (Lexerød and Eid, 2006b). Varying the scale of observation may  
102 therefore distort the information retrieved from an indicator (Chen and Crawford, 2012;  
103 Mauro et al, 2016). As plot size increases, *GC* estimates may be more reliable, but also  
104 fundamentally different stand conditions may aggregate (Coomes and Allen, 2007).  
105 Therefore, interpretation of data analysed at different scales remains one of the most  
106 challenging tasks in spatial statistics (Gotway and Young, 2002), as shown in the context of  
107 agriculture (Smith, 1938), sociology (Hannan, 1971), and environmental sciences (Jelinski  
108 and Wu, 1996). Also, the spatial distribution of trees has a practical effect on plot size, since  
109 clustered patterns require larger plot sizes to obtain a same degree of confidence in estimates  
110 (Upton and Fingleton, 1985; Pommerening, 2002; Kallimanis et al., 2008; Motz et al., 2010).  
111 Recently, Magnussen et al. (2016) suggested a method of upscaling to a common plot size to  
112 minimize scale effects in survey estimates, which achieved consistency among the quantiles  
113 and proportions of sampling distributions of forest attributes.

### 114 *1.3 Influence of ALS Scan Density in Measurements of Forest Structure*

115 Airborne laser scanning (ALS) is recognised as a highly effective tool for regional  
116 monitoring because it provides precise information about biophysical stand properties,  
117 (Gobakken et al., 2006; Gobakken and Næsset, 2008). The *GC* may be calculated as a  
118 descriptor of the distribution of ALS heights (Valbuena et al., 2017), or ALS metrics may be  
119 related to *GC* of tree sizes (Valbuena et al., 2013b). The spatial resolution of ALS data used  
120 in area-based methods has an effect on estimated values (Mascaro et al., 2011). In the context  
121 of remote sensing-assisted forest estimations, spatial resolution refers not only to the size of  
122 field plots but also to the size of pixels at which auxiliary variables are computed (Gobakken

123 and Næsset, 2008; Ruiz et al., 2014; Valbuena et al., 2016). In ALS-assisted estimations of  
124 *GC* of tree size inequality, there is a lack of knowledge on the effects of varying plot size and  
125 spatial resolution.

126 Scan density is one of the most important aspects of ALS datasets that affects both processing  
127 and costs (Balsa-Barreiro and Lerma, 2014; Kandare et al., 2016). The importance of  
128 optimizing ALS point density lays in its trade-offs against ALS swath width, and hence costs  
129 (Baltsavias, 1999). Liu et al. (2007) observed that density reduction influenced the accuracy  
130 of digital terrain models (DTM) due to the presence of empty space intervals between points.  
131 A reduction in DTM accuracy may affect the calculation of metrics describing ALS height  
132 (Ruiz et al., 2014; Singh et al., 2015), although it would be unlikely to affect metrics  
133 describing their dispersion, such as *GC*. Gobakken and Næsset (2008) assessed the effect of  
134 point density on biophysical stand properties, finding that maximum height was the least  
135 affected metric and suggesting to avoid metrics most affected by point density. No previous  
136 studies have yet determined how stand density and ALS scan densities affects *GC* estimates  
137 from ALS.

#### 138 *1.4 Objectives*

139 The aim of the study is to evaluate the effects of plot size and ALS scan density on field and  
140 ALS-derived estimates of *GC* in the boreal forests of Finland. We developed a simple method  
141 for selecting the optimal plot size for determining the *GC* of tree size inequality from field  
142 data, and for predicting *GC* reliably using ALS metrics as auxiliary variables.

## 143 **2. Material and Methods**

### 144 *2.1 Study Area and Field Data Collection*

145 The study was carried out in a typical boreal managed forest located in Eastern Finland (62°  
146 31' N, 30° 10' E). Scots Pine (*Pinus sylvestris* L.) is the dominant species which represents  
147 73% of the total wood volume, while Norway spruce (*Picea abies* Karst.) represents 16%,  
148 and deciduous species 11% of the total wood volume (Valbuena et al., 2014). The main  
149 properties of the field data such as stand density ( $N$ ), basal area ( $G$ ) and quadratic mean  
150 diameter ( $QMD$ ) are shown in **Table 1**. The field data were collected in May-June 2010 and  
151 consisted of 79 squared plots (henceforth “*original field plots*”) of various dimensions  
152 (20×20, 25×25 or 30×30 m, the smaller ones being in denser stands). With the intention of  
153 representing the contrast between highly homogeneous even-aged areas and more  
154 heterogeneous forest structures (Valbuena et al. 2016), forest stands were determined using  
155 stratified random sampling, whereas plot locations were purposively selected. After choosing  
156 the sampled stands, plots were located within the stands at a representative location. The  
157 reason for doing this was to avoid plot locations at stand borders and the high cost and  
158 measuring effort required to record the location of all individual stems within the plot. The  
159 absolute positions of every individual tree with  $dbh > 4$  cm and tree top height taller than 4 m  
160 were mapped using an approach combining ALS and field surveying methods suggested by  
161 Korpela et al. (2007). Before the field measurement, a map of individual tree positions was  
162 generated from high density ALS data (see below) using an individual tree detection (ITD)  
163 method (Packalen et al., 2013). Actual positions of trees defined by their longitude/latitude  
164 coordinates ( $X_{actual}, Y_{actual}$ ) were verified in the field, while the location of trees not  
165 detected by the ITD method were measured relative to the ITD-derived ones (distances and  
166 bearings) and least-square adjusted (Korpela et al., 2007).

167 \*\*\*approximate position of Table 1\*\*\*\*

## 168 2.2 Simulation of Circular Plots



169 Preliminary tasks for the simulation included transformations into relative coordinates, the  
170 correction of edge effects and a sensitivity analysis to determine the number of simulations  
171 needed. Then, within each *original field plot* we simulated circular plots at random positions.  
172 Circular plots were chosen on the assumption that tree competition is the same in all spatial  
173 directions. The radius of these *circular simulated plots* was increased in 1-m intervals,  
174 generating concentric circles up to 15 m-radius. Since the position of individual trees were  
175 available from the original field data, we could extract the trees located within each circular  
176 simulated plots, computing an estimation of *GC* based on tree *dbh*. Likewise, the position of  
177 individual ALS returns located within each simulated circular plots could be extracted, using  
178 them to compute ALS metrics commonly employed in area-based estimation methods.

### 179 2.2.1 Transformation to Relative Distances and Edge Correction

180 Transformation of absolute tree coordinates into relative coordinates requires procedures of  
181 plot rotation and translation (Matos, 2014). Since in the case of our study the edges of  
182 original field plots were coincident with the UTM grid, there was no need to carry out plot  
183 rotations. In plot translation absolute coordinates of original field plots were modified into  
184 relative distances, by assigning the origin of axes (0, 0) to the south-western corner of the  
185 original field plot. Absolute coordinates of south-western corner ( $X_{corner}, Y_{corner}$ ) were  
186 subtracted from the absolute coordinates of each tree ( $X_{abs}, Y_{abs}$ ) to get their relative  
187 coordinates ( $X_{rel}, Y_{rel}$ ).

$$188 \quad (X_{rel}, Y_{rel}) = (X_{abs}, Y_{abs}) - (X_{corner}, Y_{corner}) \quad (2)$$

189 Moreover, Pommerening and Stoyan (2006) showed that edge effects play an important role  
190 in spatial statistics. Because the immediate neighbour trees outside the boundary of the  
191 *original field plots* were not measured, ignoring them would result in biased statistical  
192 estimations. Thus, indices based on tree positions require an edge correction method to

193 reduce this bias. We chose a periodic boundary edge correction method (Diggle, 2003), since  
194 Pommerening and Stoyan, (2006) found it to be superior to other alternatives. This method  
195 consisted of replicating the same spatial pattern measured in the field around the *original*  
196 *field plot* (**Fig. 1**). Concentric *circular simulated plots* randomly positioned at the edge of the  
197 *original field plots* therefore also included the trees positioned out of the boundaries of the  
198 *original field plots*.

199 \*\*\*approximate position of Figure 1\*\*\*\*

### 200 2.2.2 Plot Simulation and Sensitivity Analysis

201 A pilot sensitivity analysis was done with the intention to identify the minimum number of  
202 simulations within an original field plot which can guaranteed a stable and robust outcome  
203 for the simulation. We selected the *original field plot* with highest *GC*, hence likely the one  
204 most sensible to changes among different simulations, and repeated the analysis for 10, 100,  
205 500, 700, 1000, 1500 and 2000 simulations. A position  $(X_{sim}, Y_{sim})$  was randomly  
206 simulated within the *original field plot*, and *GC* was calculated for each *circular simulated*  
207 *plot* (see below) and for each plot radius ( $s$ ; m) (1-m intervals from 1 to 15 m) (**Fig. 1**). As  
208 explained below, the standard error of the mean (*SEM*) of values obtained for *GC* at each  
209 radius were considered in order to fix the minimum number of simulations at which no  
210 considerable improvement was observed by adding further replications. After setting the  
211 necessary number of simulations to a fixed number  $k$  based on the pilot sensitivity analysis,  
212 the whole procedure was repeated for the remaining 78 original field plots. Relative and  
213 absolute positions of all simulations were recorded so that they could later be used for  
214 extracting their corresponding ALS returns as well.

### 215 2.3 Gini Coefficient Estimation

216 The target was to calculate sample estimations of the  $GC$  describing the size inequality of the  
 217 tree community represented at each *original field plot*. Its estimation (**Eq. 1**) was repeated for  
 218 every concentric *circular simulated plot* of radii 1-15 m, and for all the simulated  
 219 positions ( $X_{sim}, Y_{sim}$ ). For this purpose, basal area ( $g; m^2$ ) was calculated for each individual  
 220 stem. Differences in  $g$  were computed for each pair of trees within each circular simulated  
 221 plot.  $GC$  is the average of absolute differences relative to their mean ( $\bar{g}$ ) (see detailed  
 222 descriptions of  $GC$  calculation in Lexerød and Eid (2006a) and Valbuena et al. (2013b)). The  
 223 reason of using  $g$  instead of  $dbh$  was to increase the influence of larger trees (Solomon and  
 224 Gove, 1999). The unbiased estimator by Glasser (1962) was employed because it is  
 225 appropriate for an estimation based on a finite number of trees  $n$  located within each circular  
 226 simulated plot (**Eq. 1**). The mean  $GC$  ( $\overline{GC}$ ) and its corresponding  $SEM$  were computed for  
 227 each radius (from 1 to 15 m), and for each of the *original field plots*.  $SEM$  is a measure for  
 228 the accuracy of those means, accounting for the variability between the samples, according to  
 229 the number of simulations  $k$  and their sample standard deviation ( $SD$ ). R statistical software  
 230 (R Development Core Team, 2016) was used for all these calculations and statistical  
 231 analyses.

232 We constructed a graph comparing  $\overline{GC}$  results for increasing plot size  $s$  for all *original field*  
 233 *plots*. The  $GC$  value at *circular simulated plots* must necessarily approximate asymptotically  
 234 to the value of  $GC$  for the entire *original field plot* as the radius of circular simulated plots  
 235 increases (Matos, 2014). For this reason, the value of  $GC$  obtained by applying equation (1) to  
 236 the original field plot was used as a reference ( $GC_{ref}$ ). In order to make all the simulated  $GC$   
 237 values directly comparable, we calculated absolute  $GC$  differences ( $GC_{diff}$ ) by subtracting  
 238 simulated  $GC$  values from the  $GC_{ref}$ :

$$239 \quad \overline{GC}_{diff} = |GC_{ref} - \overline{GC}| \quad (3)$$

240 This way, it was possible to analyse the difference of each simulated *GC* to its corresponding  
241 asymptotic value, allowing to set a common criterion to evaluate all simulations based on the  
242 stabilization of the estimated *GC* value (see below).

#### 243 2.4 Airborne Laser Scanning Data and Metric Computation

244 ALS data was acquired on June 26, 2009 using ATM Gemini sensor (Optech, Canada) from  
245 600-700 m above ground level with a 26° field of view. Scan swath was 320 m wide with a  
246 55% side overlap between the strips. A high resolution dataset with 11.9 pulses·m<sup>-2</sup> scan  
247 density was produced from a pulse rate of 125 kHz. Details about the processing of ALS data  
248 are described in Packalen et al. (2013). The last echoes were classified as ground and  
249 interpolated into a DTM (Axelsson, 2000). The spatial resolution of DTM was 0.5 m based  
250 on the scan density, and the height above ground of individual ALS returns was obtained by  
251 subtraction of the DTM height beneath each of them. Echoes lower than 0.1 m from ground  
252 level were eliminated, as they were considered to be reflected from ground.

253 Individual ALS returns of each circular simulated plot based on its absolute  
254 coordinates ( $X_{sim}, Y_{sim}$ ) were clipped, and area-based ALS metrics were computed from their  
255 heights with the help of FUSION software (USDA Forest Service; McGaughey, 2015). ALS  
256 metrics are statistics and descriptors of the distribution of ALS heights observed within a  
257 given area, which are usually employed as auxiliary variables in ALS-assisted forest  
258 estimations (**Table 2**). Some of these metrics were common statistics as, for example, the  
259 mean (*Mean*) standard deviation (*StdDev*) or the skewness (*Skew*) of the distribution of  
260 heights above ground of ALS returns contained within each *circular simulated plot*. We also  
261 computed the percentiles of their distribution, such as the 25<sup>th</sup> (*P25*), 50<sup>th</sup> (*P50*) or 99<sup>th</sup> (*P99*).  
262 In addition, we calculated the so-called canopy cover metrics (McGaughey, 2015), such as  
263 the proportion of returns backscattered from 0.1 m above the ground (*Cover*). Another metric

264 included in FUSION was the canopy relief ratio (*CRR*), which is the difference between  
265 mean and minimum ALS return heights divided by a difference between maximum and  
266 minimum heights (Pike and Wilson, 1971).

267 \*\*\**approximate position of Table 2*\*\*\*\*

268 The effect of plot size in the relationship with *GC* was studied separately for each of these  
269 ALS metrics. For each radius, we gathered all the simulations carried out at all the *original*  
270 *field plots* and calculated all the ALS metrics listed in **Table 2**. They were used to calculate  
271 Pearson correlation coefficients (*r*) using all the combinations of field *GC* against each ALS  
272 metric. Then, we observed separately for each ALS metric the evolution of *r* when increasing  
273 the plot size *s* of the *circular simulated plots*. Since we were only interested in the capacity of  
274 the ALS metrics to explain variability in *GC*, regardless of whether their relationship was  
275 direct or indirect, we considered the absolute value of the correlation coefficient  $|r|$  in the  
276 optimization, as explained below.

### 277 *2.5 Basic Relationships*

278 The plot size and spatial resolution at which an ALS-assisted estimation is carried out relates  
279 intrinsically to the sample size used in all calculations. Sample size affects the relationship  
280 between predictor and response, and therefore the accuracy of ALS estimation of any forest  
281 attributes (Gotway and Young, 2002; Mascaro et al., 2010; Næsset et al., 2015; Magnussen et  
282 al., 2016; Valbuena et al., 2016). In this context, sample size refers both to the number of  
283 trees used to calculate a given forest attribute, *GC* in this case, but also to the number of ALS  
284 returns involved in the computation of ALS metrics. The link between resolution and sample  
285 size is employed on the empirical densities of the datasets, i.e. stand density ( $N$ ; trees·ha<sup>-1</sup>) or  
286 ALS points density ( $d$ ; points·m<sup>-2</sup>) (Gobakken and Næsset, 2008; Motz et al., 2010;  
287 Jakubowski et al., 2013). Therefore, the effects of plot size and spatial resolution of the ALS

288 estimated forest attributes also depend on  $N$  and  $d$ , and the combined effects of these two  
289 factors may explain why plot sizes suitable for field surveys may be found sub-optimal for  
290 ALS estimation (Næsset et al., 2015).

291 Hence, the relationship between the radius  $s$  of a circular plot and the number of trees ( $n$ )  
292 contained within is tied to the  $N$  at the location of the plot.

$$293 \qquad n = N\pi s^2 \qquad (4)$$

294 This begs the question on whether the optimization method should search for an optimal plot  
295 radius ( $s^*$ ; m) or an optimal sample size ( $n^*$ ). In a forest environment of variable stand  
296 density  $N$  (**Table 1**), does the relationship between  $GC$  and ALS metrics depends on the plot  
297 size used, or on the number of trees surveyed? In order to research whether it makes a  
298 difference, we repeated the same procedure for both  $s^*$  and  $n^*$  optimization. In other words,  
299 we tested the results of optimization according to either plot radius or number of trees. In any  
300 of the cases, the relationship in **eq. (4)** assures that the methodology can be replicated for  
301 either dense or sparse forests, since  $s$  and  $n$  can always be deduced from one another by an  
302 empirical  $N$ .

303 Likewise, a similar relationship holds between the size of that same circular plot and the  
304 number of ALS returns backscattered from it, according to a given ALS scan density  $d$ . In  
305 this context of estimation using auxiliary variables, the scale concerns both to the size of the  
306 field plots and the spatial resolution of the pixel at which ALS metrics are calculated.  
307 Therefore, the number of ALS points ( $p$ ) relates to the spatial resolution / plot size used ( $s$ )  
308 according to  $d$  :

$$309 \qquad p = d\pi s^2 \qquad (5)$$

310 As before, the relationship in eq. (5) assures that the methodology can be replicated for any  
311 range of ALS scan densities, since  $s$  and  $p$  are trivially deducted from one another by an  
312 empirical  $d$ . As an overall conclusion, a given optimal plot size  $s^*$  necessarily implies  
313 optimal sample sizes as well, both  $n^*$  and  $p^*$ . Keeping these relationships in mind is key to  
314 demonstrating the validity of the optimization method for replication elsewhere according to  
315 the  $N$  and  $d$  which may occur at any other study cases, and therefore the method is equally  
316 valid for both dense and sparse forests and ALS surveys with low or high scan density.

### 317 *2.6 Plot Size Optimization*

318 To optimize the plot size which should be used for a reliable  $GC$  estimation, and thereby also  
319 the optimal spatial resolution for an estimation of  $GC$  from ALS datasets, we determined two  
320 criteria to be applied sequentially: (1) stabilization of  $GC$  as estimator of the population value  
321 from the field information itself, and (2) maximizing the  $GC$  variability explained by ALS  
322 metrics. Therefore, *Criterion I* considered the minimum plot radius at which the estimation of  
323  $GC$  remained stable to further increases in plot size. *Criterion II* was set to optimize the ALS-  
324 assisted estimation, by observing changes in the correlation between the field  $GC$  and each  
325 ALS metric among the simulated plot radii.

326 *Criterion I* was implemented by observing the evolution of  $\overline{GC}_{diff}$  for increasing radii at  
327 every original field plot. We set a maximum value of  $\overline{GC}_{diff} = 0.05$  at which it was  
328 considered that the estimation of  $GC$  was stable and representative of the population, and,  
329 therefore, selected the minimum plot radius  $s$  as the smallest meeting the first criterion for all  
330 the 79 original field plots.

331 *Criterion II* consisted in maximizing the explained variance in the  $GC$  values when predicted  
332 from ALS metrics. To implement this criterion we combined all the  $GC/ALS$  metric pairs for

333 all the simulations carried out at all the original field plots, and grouped them according to  
 334 the different simulated radii. The optimal radius was set to be that one showing the maximum  
 335  $|r|$  value for a given metric. To make an overall decision, we put the focus on those metrics  
 336 showing higher correlations, and decided a range of optimal sizes accordingly (since the  
 337 empirical maximum may differ for different ALS metrics). As a summary, the final optimal  
 338 plot size  $s^*$  for a given metrics was:

$$339 \quad s^* = \max|r| \mid \overline{GC}_{diff} < 0.05 \quad (6)$$

### 340 *2.7 Sample Size Optimization*

341 For sample size optimization, seeking to deduct what is the minimum number of trees needed  
 342 to obtain a reliable  $GC$  estimation, and the optimum for its ALS prediction, we applied the  
 343 same two sequential criteria employed for plot size optimization (section 2.6). Therefore, the  
 344 simulations were similar as before, but they increased the size of simulated circular plots  
 345 according to the resulting number of trees  $n$  instead of plot radii. Thus, for implementing  
 346 *Criterion I*, the evolution of  $\overline{GC}_{diff}$  was observed for increasing number of trees  $n$ , also  
 347 setting a maximum value of  $\overline{GC}_{diff} = 0.05$ . As before, we selected the minimum  $n$  as the  
 348 smallest meeting *Criterion I* for all 79 original field plots. *Criterion II* also consisted in  
 349 maximizing the absolute correlation between the  $GC$  values and each of the ALS metrics.  
 350 New values of  $|r|$  were obtained for increasing values of  $n$ , and the final optimal sample size  
 351 ( $n^*$ ) for each given ALS metric was then set as:

$$352 \quad n^* = \max|r| \mid \overline{GC}_{diff} < 0.05 \quad (7)$$

353 Finally, we compared which alternative, **eq. (6)** or **(7)**, would be more convenient for a  
 354 practical plot size optimization, discussing the results obtained by either method.



## 355 2.8 Reduction of ALS Point Density

356 Once deduced an optimal spatial resolution  $s^*$ , we also investigated the effects of varying  
357 ALS scan density  $d$ . The original point density was reduced to 0.5, 0.75, 1, 3, 5, 7.5, and 10  
358 points·m<sup>-2</sup>. A common option to reduce point density is by moving a 1 m window and  
359 selecting random points from the point cloud to reach the desired point density (e.g.,  
360 Magnussen et al. 2010). We calculated a correct thinning factor for each desired point density  
361  $d$  (Ruiz et al., 2014), following the method detailed by Jakubowski et al. (2013) which  
362 incorporates routines included in *LAStools* (RapidLasso GmbH Inc.; Isenburg, 2016). New  
363 ALS metrics over each of the  $k$  simulated circular plot positions and their correlations against  
364 the  $GC$  values obtained from the field information were calculated, and the entire procedure  
365 was repeated for all the reduced densities. In a similar manner as it was done for  $s$  and  $n$ , the  
366 effects of varying ALS scan density were studied by observing the changes in  $|r|$ , i.e. the  
367 effects in the relationship between the  $GC$  of tree size inequality and the ALS metrics with  
368 more explanatory capacity towards this given forest attribute.

## 369 3. Results

### 370 3.1 Establishing the Number of Simulations

371 **Figure 2** shows the results of sensitivity analysis carried out to select the minimum number  
372 of simulations that would yield a robust estimation of  $GC$  for increasing simulated plot radii.  
373 As expected, the  $GC$  value estimated from few simulations fluctuated considerably, and this  
374 fluctuation decreased as the number of simulations increased (**Fig. 2a**). The expected general  
375 trend toward the asymptotic value obtained by the entire population ( $GC_{ref}$ ) was generally  
376 observed in **Fig. 2a**. Very little variation in  $GC$  estimates were observed when the number of  
377 simulations increased from 700. Similarly, the  $SEM$  decreased as the number of simulations  
378 increased (**Fig. 2b**), remaining virtually unchanged from 700 to 2000 simulations.

379 Consequently, we decided to carry out the analysis using  $k = 700$  simulations of 15  
380 concentric circular simulated plots located within each 79 original field plots.

381 \*\*\*approximate position of Figure 2\*\*\*\*

### 382 3.2 Plot Size Optimization

383 **Figure 3a** shows the resulting  $\overline{GC}_{diff}$  for each of the 79 original plots, and **Table 3** is a  
384 summary of these results which was used for establishing *Criterion I*, which set the minimum  
385 plot size that would provide a reliable *GC* estimation for the population. Circular simulated  
386 plots of small sizes provided *GC* estimates that differed considerably from the population  
387 values as considered by  $GC_{ref}$ . Nonetheless, once the estimation reached stabilization, an  
388 increase in the radius of a circular plot (and hence the sampling effort) would not necessarily  
389 imply a considerable change in the estimation of *GC* (**Fig. 3a**). Our results showed that only  
390 few of the original field plots (probably very homogeneous stands) obtained stable *GC*  
391 estimations from very small circular simulated plots (**Table 3**). On the other hand, for larger  
392 circular simulated plots the differences against the original field plots representing the  
393 population became negligible. We observed that stabilization of the *GC* estimation started  
394 beyond of simulated plot radius  $s = 6$  m, from which all the original field plots fell within  
395 the  $\overline{GC}_{diff} < 0.05$  limit. We therefore established that the smallest plot size required for a  
396 reliable *GC* estimation should be set at areas sizing around 113 m<sup>2</sup>.

397 \*\*\*approximate position of Figure 3\*\*\*\*

398 \*\*\*approximate position of Table 3\*\*\*\*

399 With regards to *Criterion II*, the evolution of  $|\tau|$  with increasing plot size was observed for  
400 all ALS metrics included in FUSION. Results showed that changes in the relationship  
401 between the field *GC* of tree sizes and metrics describing the distribution of ALS return

402 followed some general trends and patterns. For this reason and for simplifying, we chose to  
403 show only few ALS metrics in **Fig. 4a**, which we considered representatives of the general  
404 trends observed. These ALS metrics were the described  $P25$ ,  $P50$ ,  $P99$ ,  $Skew$ ,  $StdDev$ ,  $Cover$   
405 and  $CRR$  (**Table 2**). **Fig. 4a** showed an erratic fluctuation for the values of  $|r|$  obtained for  
406 plot sizes smaller than a radius  $s = 5$  m, which was possibly caused by the instability  
407 observed in the  $GC$  estimation at smaller plot sizes (**Fig. 3**). For this reason, we shadowed this  
408 area in grey colour in **Fig. 4**, denoting that such small plot sizes were already dismissed under  
409 *Criterion I*. Once  $GC$  estimation reached stabilization, its correlation to ALS metrics often  
410 yielded a convex curve as plot size increased (**Fig. 4a**). Therefore, the optimal plot size was  
411 possible to determine via maximization of  $|r|$ . This tendency was more clearly marked for  
412 those ALS metrics showing higher values of  $|r|$ , i.e. more correlated to the  $GC$  of tree sizes  
413 (**eq. 2**), such as  $Skew$ ,  $Cover$  or  $CRR$ . For other ALS metrics less related to  $GC$ , like return  
414 height percentiles ( $P25$ ,  $P50$  or  $P99$ ) or  $StdDev$ , this tendency was less marked (**Fig. 4a**). For  
415 the optimization of plot size, we selected those metrics showing highest correlation  
416 against  $GC$ , since in practice they would be those more involved in its estimation. **Table 4**  
417 shows that the maximum  $|r|$  for ALS metrics  $Skew$ ,  $Cover$  or  $CRR$  ranged  $s^* = 9-12$  m plot  
418 radius (the quality of histograms and scatterplots between variables involved can be checked  
419 in the **Supplementary Material**). It can be observed in **Fig. 4a** that beyond a circular  
420 simulated plot of 12 m the correlation showed a decreasing trend for most ALS metrics. Also,  
421 local maxima may be found for some ALS metrics for very small plot sizes, which is  
422 probably an artefact due to the above-mention instability in  $GC$  estimation at very small plot  
423 sizes (**Fig. 3**). This proved the necessity of imposing *Criterion I* as a prior step to correlation  
424 maximization. As a conclusion, under the established combined *Criteria I* and *II*, we  
425 determined that any plot radius  $s < 6$  m (113 m<sup>2</sup> area) should be avoided (denoted by grey  
426 colour in **Fig. 4a**), and the optimal plot size for an ALS-assisted estimation of  $GC$  must be

427 carried out using scales sizing 250-450 m<sup>2</sup>, which concerns to both the size of the field plot  
428 and the pixel of the grid employed for ALS estimation.

429 \*\*\*approximate position of Figure 4\*\*\*\*

430 \*\*\*approximate position of Table 4\*\*\*\*

### 431 3.3 Sample Size Optimization (Stand Density Effect)

432 On the other hand, **Figure 3b** shows the evolution of  $\overline{GC}_{diff}$  for increasing sample sizes  
433 (number of trees  $n$ ) at each of the 79 original field plots. It is worth mentioning the **Figs. 3a**  
434 and **3b** relate to one another according to **eq. (4)**. As a consequence, a similar tendency can  
435 be found for both of them. **Table 3** expresses the number of trees that correspond on average  
436 to a given sample size. Therefore, the minimum value obtained for *Criterion I* in plot size  
437 optimization,  $s = 6$ , corresponds to stating that a minimum number of  $n = 15$  trees are  
438 required for a stable  $GC$  estimation (shaded area in **Fig. 4b**). We nevertheless further  
439 postulated that this minimum number of trees may be dependent on the heterogeneity of the  
440 forest itself, being possibly larger in the presence of higher inequality of tree sizes. This  
441 presumption was demonstrably true, as it can be observed in a scatterplot comparing the  
442 minimum number of trees required for a stable  $GC$  estimation at each of the 79 original plots  
443 against their overall value of tree size inequality observed ( $GC_{ref}$ ; **Fig. 5**). Such relationship  
444 was not so straightforward if *Criterion I* was imposed using  $s$  instead (results not shown),  
445 which demonstrates the effect of varying forest stand density  $N$ . Hence, obtaining a stable  $GC$   
446 estimation is more dependent of measuring a minimum number of trees than imposing a  
447 given size for the field plot used.

448 \*\*\*approximate position of Figure 5\*\*\*\*

449 The case for *Criterion II* was different, as it can be deducted when observing the same ALS  
450 metrics employed to optimize  $s$  –  $P25$ ,  $P50$ ,  $P99$ ,  $Skew$ ,  $StdDev$ ,  $Cover$  and  $CRR$  –, but trying  
451 to optimize  $n$  instead (**Fig. 4b**). Again, a similar tendency can be found since **Figs.4a-b** are  
452 also related by **eq. (4)**. Results were therefore very similar whether optimization was carried  
453 out according to plot size (**eq. 6**) or sample size (**eq. 7**). The values of  $|r|$  also followed a  
454 convex curve when increasing the number of trees measured, and an optimal sample size  $n^*$   
455 could be reliably determined via  $|r|$  maximization. Our results showed that a number of trees  
456 approximately ranging  $n^* = 30-60$  (**Table 4**) should be involved in the computation of  $GC$ ,  
457 in order to maximize the efficiency of its estimation using ALS. Since the value of  $|r|$   
458 involves both the field  $GC$  and the ALS metrics, its changes are determined by both  $N$  and  $d$   
459 (**eqs. 4-5**), and both may cause a change in the correlation between the two variables.

#### 460 3.4 Effect of Point Density on the Relationship of $GC$

461 According to the previous results, we set the optimal plot size to  $s^* = 9$  m in order to further  
462 analyse the possible effects due to varying scan density. Among all the ALS metrics (**Table**  
463 **2**), we selected those same ones employed previously –  $P25$ ,  $P50$ ,  $P99$ ,  $Skew$ ,  $StdDev$ ,  $Cover$   
464 and  $CRR$  – to allow direct comparison. **Fig. 6** shows the evolution in  $|r|$  for increasing ALS  
465 point density  $d$ . No considerable changes were observed in the correlation between the field  
466  $GC$  and the ALS metrics, which suggests that  $d$  has no major effects on their relationship.  
467 However, a decreasing trend in  $|r|$  could generally be observed when point densities  
468 decreased below  $d < 3$  points·m<sup>-2</sup> (**Fig. 6**). Overall, these results therefore suggest that the  
469 relationship between  $GC$  and ALS metrics is mainly dependent on the plot size employed,  
470 and rather independent of stand density and ALS scan density

471 \*\*\*approximate position of Figure 6\*\*\*\*

472 **4. Discussion**

473 In this study we evaluated the effects of plot size and sample size on the  $GC$  of tree size  
474 inequality, and on its practical estimation using remote sensing methods based on ALS.  
475 Sample size refers to the number of individual elements (trees or ALS returns) included  
476 within a given sample area, which is therefore determined by the spatial resolution employed  
477 for evaluating a given forest attribute. We also analysed the effects of ALS scan density and,  
478 overall, we observed that plot size had greater effects on the relationship between  $GC$  and  
479 ALS metrics than either of the other two criteria considered. The motivation for studying  
480 these effects is grounded on the fact that inappropriate plot sizes may provide unreliable  
481 estimates and lead to sub-optimal forest management decisions (Eid, 2000; Mauro et al.,  
482 2010). Valbuena et al. (2013a) pointed out that the estimation of  $GC$  is affected by the area at  
483 which it is evaluated. Results in **Fig. 3** illustrate how the  $\overline{GC}_{diff}$  decreases when increasing  
484 the size of circular plots and, and hence their corresponding sample size.  $\overline{GC}_{diff}$  values  
485 markedly dropped for smaller plot radii and sample sizes. This decrease smooths from bigger  
486 sizes, which indicates stabilization of the estimation (*Criterion I*). **Fig. 2a** also shows an  
487 example of this tendency to asymptotically approach the population value, which was also  
488 observed by George (2003), Barbeito et al. (2009), or Matos (2014). Based on *Criterion I*  
489 ( $\overline{GC}_{diff} < 0.05$ ), the circular plot should be large enough ( $s \geq 6$  m) to have minimum  
490 sample size of  $n \geq 15$  trees (**Fig. 3**). Although the minimum plot size also depends on the  
491 stand density of an area, **eq. (4)** can be used to adjust the method to any forest areas, whether  
492 sparsely or densely forested. This conclusion may therefore be partly extended to other forest  
493 types, as it can be for example deduced (via **eq. 4**) that minimum radius of  $s \geq 12$  m would  
494 be needed in sparsely forested area of only  $300 \text{ stems} \cdot \text{ha}^{-1}$  (Lombardi et al., 2015). **Eq. (4)**  
495 therefore brings generality to the method, since plot sizes may hence be tailored to forest  
496 areas of differing stand densities.

497 In this article we also postulated that maximizing the explained variability between the *GC*  
498 estimated from the field and ALS metrics could be a valid criterion to optimize the reliability  
499 of ALS-assisted estimations of *GC* (*Criterion II*). Results in **Fig. 4a** showed that our  
500 presumption was correct, since the  $|r|$  values between *GC* and most ALS metrics, especially  
501 the most correlated ones, followed a convex curve with a maximum that could be searched to  
502 reach an optimal plot size / spatial resolution for the estimation. On the other hand, once the  
503 *GC* reached some stabilization, the correlation between them remains largely unchanged.  
504 Therefore, a lower plot size limit is to be imposed to avoid local minima that could appear as  
505 an artefact of the unstable estimation of *GC* at low sample sizes. We shaded this area in grey  
506 colour in **Fig. 4 (a, b)**, denoting the area that was already dismissed as a result of *Criterion I*  
507 (**Fig. 3**; George, 2003). In larger plots the sample size was more representative of the total  
508 population. Combining both criteria, we found in our study area that an optimal circular plot  
509 radius of  $s^* = 9\text{-}12$  m, which corresponds to a spatial resolution of sampling units sizing  
510  $250\text{-}450$  m<sup>2</sup> (**Fig. 4a**), would be suitable for ALS-assisted *GC* estimation. Since plot size and  
511 sample size are interdependent (**eq. 4**), this result may be suitable for any area with a similar  
512 average number of trees ( $N \cong 1300$  stems·ha<sup>-1</sup>; **Table 2**). According to these results,  
513 therefore, most forest datasets commonly acquired in operational inventories would be  
514 acceptable for an ALS-assisted estimation of the *GC* of tree sizes. Lombardi et al. (2015)  
515 deduced a larger optimal plot radius  $s^* = 13\text{-}15$  m for other forest attributes, most likely due  
516 to lower  $N$  in the forest areas considered. For studies dealing with differing plot sizes, one  
517 possibility could be to upscale *GC* to a common plot size (Kent and Coker, 1992; Magnussen  
518 et al., 2016).

519 Some of the reflexions raised in this article affect all other types of forest attributes and  
520 remotely sensed auxiliary variables that may be used in forest estimations (Jelinski and Wu,  
521 1996). However, different forest attributes are differently affected by varying plot sizes

522 (Chytrý and Otýpková, 2003). Some forest variables such as stand density or biomass would  
523 show an averaging effect as plot size increases (Jelinski and Wu, 1996; Gotway and Young,  
524 2002; Ruiz et al., 2014), which in turn derives in improved model efficiency when using  
525 larger scales in remote sensing estimations (Næsset et al., 2015; Mauro et al., 2016). But  
526 there is a trade-off between model accuracy and spatial resolution, and root mean squared  
527 errors increase from 10-15% for 1000-4000 m<sup>2</sup> to 20-25% for 200-250 m<sup>2</sup> (Næsset, 2002,  
528 2004, 2007). However, this averaging effect is not applicable to forest attributes describing  
529 structural diversity and heterogeneity (Coomes and Allen, 2007). In fact, many variables  
530 necessarily augment when the plot size increases, for instance species richness and diversity  
531 (e.g., Humphrey et al. 2000; Otypková and Chytry, 2006; Kallimanis et al., 2008; Fibich et  
532 al., 2016) as traditionally assessed through rarefaction (Kent and Coker, 1992). A similar  
533 effect can be observed in other measures of forest heterogeneity (Barbeito et al., 2009; Motz  
534 et al., 2010; McRoberts et al., 2012), and thus in the *GC* (Valbuena et al., 2013a, Matos,  
535 2014), since increasing the size of a plot increases the probability of finding an additional  
536 differently-sized tree (Chen and Crawford, 2012; Valbuena et al., 2012). This is why  
537 estimated *GC* values in **Fig. 3** asymptotically approach the value of the larger original field  
538 plot (George, 2003; Matos, 2014), which is never exceeded. Imposing a criterion defining  
539 which of the plausible plot sizes should be used is therefore not a trivial question to tackle.  
540 Matos (2014) employed a number of different criteria based on the field information only –  
541 stabilization of the estimate, stabilization of certainty of the estimate and convergence with  
542  $GC_{ref}$  –, none of them resulting fully satisfactory and definitive as they all ultimately rely on  
543 a subjective assumption (Cressie, 1993). For this reason, in this article we approached the  
544 question of plot size from the viewpoint of its practical estimation using ALS remote sensing.  
545 The convex curves obtained in **Fig. 4a** proved this approach to be highly beneficial, since  
546 maximization of correlation  $|r|$  between *GC* values and selected ALS predictors provides



547 with a more objective method for determining the optimal plot size for the assessment of  $GC$   
548 of tree size inequality. Still, due to the very high uncertainty observed in the estimation of  $GC$   
549 when using very small plot sizes (**Fig. 3b**; Smith, 1938; Lombardi et al., 2015), we deduced  
550 that a criterion avoiding great divergence with  $GC_{ref}$  may be imposed as a prior step to  
551 maximization (Motz et al. (2010) referred to it as minimum grid spacing). Further research  
552 could focus on modelling  $GC$  from ALS metrics and investigate how the interaction among  
553 many ALS metrics in a same model may play a relevant role in the optimization of plot size  
554 and spatial resolution.

555 The analyses carried out with reduced point densities revealed that lowering point density  
556 barely affects the correlation between  $GC$  and ALS metrics, unless using a very sparse scan  
557 density  $d < 3$  points·m<sup>-2</sup>. Previous studies such as Maltamo et al. (2006), Ruiz et al. (2014)  
558 or Singh et al. (2015) also indicate that reducing the point density is not affecting the  
559 accuracy of volume prediction and demonstrate that the effects of varying scan densities can  
560 be eluded in practical applications. It must be taken into account, however, that the DTM  
561 used in this study was based on original point density, and the errors in DTM determination  
562 at sparser densities (Liu et al., 2007; Ruiz et al., 2014) may induce to further uncertainty,  
563 although this presumably has a lesser effect on those metrics most related to  $GC$ .  
564 Furthermore, since ALS datasets from national programmes are currently surveying entire  
565 countries at densities typically between 0.5-1 points·m<sup>-2</sup> (Artuso et al., 2003), it must also be  
566 pointed out the relevance of results in **Fig. 6** which render most of these nation-wide ALS  
567 datasets unsuitable for reliably estimating  $GC$  (Kandare et al. 2016). In line with results in  
568 Valbuena et al. (2017), who postulated that the low densities incur in critical omission of  
569 understorey development, our results demonstrate that indeed there is a need for increasing  
570 point densities up to  $d = 3$  points·m<sup>-2</sup>. This result is very concurrent with those obtained by  
571 Ruiz et al. (2014) and Watt et al. (2014) for different forest attributes in different stand types,

572 and therefore the case seems clear that ALS datasets obtained for forest applications should  
573 reach this minimum density requirement.

## 574 **5. Conclusion**

575 In this study we studied how changing spatial resolution can affect the relationship between  
576 *GC* and ALS metrics. We used three criteria for optimization: plot size, stand density and  
577 ALS scan density. The effects of stand and scan densities are intimately interrelated to plot  
578 size, since they together determine the sample size employed in calculations. Amongst those  
579 three criteria, we found plot size to predominantly affect the relationship between *GC* and  
580 ALS metrics.

581 We observed that the estimation of *GC* is strongly affected by the size of the forest plot  
582 surveyed. Very small sample size and plot radii are more sensitive to *GC* variations,  
583 unrepresentative of the total population, producing unstable and unreliable *GC* estimations.  
584 The *GC* estimation stabilizes as the size of plots and samples increases, as larger plots contain  
585 a more appropriate number of observations (sample size) representing the population. We  
586 determined that, in a boreal managed forest, a minimum number of 15 trees ought to be  
587 measured for a reliable *GC* estimation, regardless of the stand density present at each forest  
588 stand.

589 We developed a method for plot size optimization based on a combination of two criteria: (1)  
590 imposing a minimum of number of 15 trees measured, and (2) maximizing the absolute  
591 correlation between field *GC* and ALS metrics. The plot level correlation between ALS  
592 metrics and field *GC* showed a convex tendency for increasing plot sizes. Our results showed  
593 that 9-12 m-radius plots produced the maximum correlation thus they are suitable for ALS-

594 assisted *GC* estimation. Basic relationships between plot size and sample size may be used to  
595 accommodate the method to forested environments of varying stand densities.

596 With regards to the effects of ALS scan density, we observed that it can barely have any  
597 effects unless lowered under 3 points  $m^{-2}$ . This however may be relevant for the practical  
598 application of low-density national datasets, and therefore we would recommend increasing  
599 their scan densities with the intention to render nation-wide datasets useful for studying forest  
600 heterogeneity.

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### 607 **References**

- 608 Asada, Y., 2005. Assessment of the health of Americans: the average health-related quality of  
609 life and its inequality across individuals and groups. *Population Health Metrics* 3, 7-7.
- 610 Artuso, R., Bovet, S., Streilein, A. 2003. Practical Methods for the Verification of  
611 Countrywide Terrain and Surface Models. ISPRS WG III/3 Workshop. 3-D reconstruction  
612 from airborne laser scanner and InSAR data. Dresden, Germany 8-10 October 2003. In: The  
613 International Archives of Photogrammetry, Remote Sensing and Spatial Information  
614 Sciences, Vol. XXXIV, part 3/WG13.
- 615 Axelsson, P., 2000. DEM generation from laser scanner data using adaptive TIN models.  
616 *International Archives of Photogrammetry and Remote Sensing* 33, 111-118.
- 617 Balsa-Barreiro, J., Lerma, J.L., 2014. Empirical study of variation in lidar point density over  
618 different land covers. *Int. J. Remote Sens.* 35, 3372-3383.
- 619 Baltsavias, E.P., 1999. Airborne laser scanning: basic relations and formulas. *ISPRS Journal*  
620 *of photogrammetry and remote sensing* 54, 199-214.

- 621 Barbeito, I., Cañellas, I., Montes, F., 2009. Evaluating the behaviour of vertical structure  
622 indices in Scots pine forests. *Ann. For. Sci.* 66, 710-710.
- 623 Bollandsås, O.M., Næsset, E., 2007. Estimating percentile-based diameter distributions in  
624 uneven-sized Norway spruce stands using airborne laser scanner data. *Scand. J. For. Res.* 22,  
625 33-47.
- 626 Bourdier T., Cordonnier T., Kunstler G., Piedallu C., Lagarrigues G., Courbaud B. 2016.  
627 Tree size inequality reduces forest productivity: an analysis combining inventory data for ten  
628 european species and a light competition model. *PLoS ONE* 11(3), e0151852.
- 629 Buongiorno, J., Dahir, S., Lu, H., Lin, C., 1994. Tree Size Diversity and Economic Returns in  
630 Uneven-Aged Forest Stands. *For. Sci.* 40, 83-103.
- 631 Chen, Z., Crawford, C.A.G., 2012. The role of geographic scale in testing the income  
632 inequality hypothesis as an explanation of health disparities. *Soc. Sci. Med.* 75, 1022-1031.
- 633 Chytrý, M., Otýpková, Z., 2003. Plot sizes used for phytosociological sampling of European  
634 vegetation. *Journal of Vegetation Science* 14, 563-570.
- 635 Coomes, D.A. and Allen, R.B., 2007. Mortality and tree-size distributions in natural mixed-  
636 age forests. *Journal of Ecology*, 95(1), pp.27-40.
- 637 Cordonnier, T., and Kunstler, G. 2015. The Gini index brings asymmetric competition to  
638 light. *Perspectives in Plant Ecology, Evolution and Systematics*, 17 (2), 107-115.
- 639 Cressie, N., 1993. Aggregation in Geostatistical Problems, in Soares, A. (Ed.), *Geostatistics*  
640 *Troia'92: Volume 1*. Springer Netherlands, Dordrecht, pp. 25-36.
- 641 David, H.A., Mishriky, R.S., 1968. Order Statistics for Discrete Populations and for Grouped  
642 Samples. *Journal of the American Statistical Association* 63, 1390-1398.
- 643 Diggle, P.J., 2003. *Statistical Analysis of Spatial Point Patterns*, 2nd ed. Arnold, London.
- 644 Duduman, G., 2009. An ecological approach for establishing the allowable cut in forests  
645 where single tree selection system is applied. Editura Universităţii Suceava.[In Romanian] .
- 646 Eid, T., 2000. Use of uncertain inventory data in forestry scenario models and consequential  
647 incorrect harvest decisions. *Silva Fenn.* 34, 89-100.
- 648 Fibich, P., Lepš, J., Novotný, V., Klimeš, P., Těšitel, J., Molem, K., Damas, K., Weiblen,  
649 G.D., 2016. Spatial patterns of tree species distribution in New Guinea primary and  
650 secondary lowland rain forest. *Journal of Vegetation Science* 27, 328-339.
- 651 George, D., 2003. The Small-Sample Bias of the Gini Coefficient: Results and Implications  
652 for Empirical Research. *Review of Economics and Statistics* 85, 226-234.
- 653 Gini, C., 1921. Measurement of inequality of incomes. *The Economic Journal* 31, 124-126.

- 654 Glasser, G.J., 1962. Variance formulas for the mean difference and coefficient of  
655 concentration. *Journal of the American Statistical Association* 57, 648-654.
- 656 Gobakken, T., Næsset, E., Nelson, R., 2006. Developing regional forest inventory procedures  
657 based on scanning LiDAR. *Proceedings of the International Conference Silvilaser 2006*,  
658 Matsuyama, Japan. Japan Society of Forest Planning, Forestry and Forest Products Research  
659 Institute and Ehime University. pp. 99-104.
- 660 Gobakken, T., Næsset, E., 2008. Assessing effects of laser point density, ground sampling  
661 intensity, and field sample plot size on biophysical stand properties derived from airborne  
662 laser scanner data. *Canadian Journal of Forest Research* 38, 1095-1109.
- 663 Gotway, C.A., Young, L.J., 2002. Combining incompatible spatial data. *Journal of the*  
664 *American Statistical Association* 97, 632-648.
- 665 Hannan, M.T., 1971. *Aggregation and Disaggregation in Sociology*. Lexington Books.
- 666 Humphrey, J., Newton, A., Peace, A., Holden, E., 2000. The importance of conifer  
667 plantations in northern Britain as a habitat for native fungi. *Biol. Conserv.* 96, 241-252.
- 668 Hvistendahl, M., 2014. While emerging economies boom, equality goes bust. *Science*, 344,  
669 832-835
- 670 Isenburg, M., 2016. "LASStools—efficient tools for LiDAR processing". (Version 160921,  
671 academic). Retrieved from <http://rapidlasso.com/LASStools>.
- 672 Jakubowski, M.K., Guo, Q., Kelly, M., 2013. Tradeoffs between lidar pulse density and  
673 forest measurement accuracy. *Remote Sens. Environ.* 130, 245-253.
- 674 Jelinski, D., Wu, J., 1996. The modifiable areal unit problem and implications for landscape  
675 ecology. *Landscape Ecol.* 11, 129-140.
- 676 Kallimanis, A.S., Halley, J.M., Vokou, D. & Sgardelis, S.P., 2008. The scale of analysis  
677 determines the spatial pattern of woody species in the Mediterranean environment. *Plant*  
678 *Ecology* 196: 143–151.
- 679 Kandare, K., Ørka, H.O., Chan, J.C.-W., Dalponte, M., 2016. Effects of forest structure and  
680 airborne laser scanning point cloud density on 3D delineation of individual tree crowns.  
681 *European Journal of Remote Sensing*, 49, pp. 337-359.
- 682 Kent, M., Coker, P., 1992. *Vegetation description and analysis: a practical approach*.  
683 Chichester: Wiley.
- 684 Knox, R.G., Peet, R.K., Christensen, N.L., 1989. Population dynamics in loblolly pine stands:  
685 changes in skewness and size inequality. *Ecology* 70, 1153-1167.
- 686 Korpela, I., Tuomola, T., Välimäki, E., 2007. Mapping forest plots: an efficient method  
687 combining photogrammetry and field triangulation. *Silva Fenn.* 41.

- 688 Lähde E., Laiho O., Norokorpi Y., 1999. Diversity-oriented silviculture in the boreal zone of  
689 Europe. *Forest Ecology and Management* 118: 223–243.
- 690 Lei X., Wang W., Peng, C., 2009. Relationships between stand growth and structural  
691 diversity in spruce-dominated forests in New Brunswick, Canada. *Canadian Journal of Forest*  
692 *Research* 39, 1835–1847.
- 693 Lexerød, N., Eid, T., 2006a. An evaluation of different diameter diversity indices based on  
694 criteria related to forest management planning. *For. Ecol. Manage.* 222, 17-28.
- 695 Lexerød, N., Eid, T., 2006b. Assessing suitability for selective cutting using a stand level  
696 index. *Forest Ecology and Management*, 237(1–3), pp. 503-512.
- 697 Liu, X., Zhang, Z., Peterson, J., Chandra, S., 2007. The effect of LiDAR data density on  
698 DEM accuracy. *Proceedings of the International Congress on Modelling and Simulation*  
699 (MODSIM07) 2007, Modelling and Simulation Society of Australia and New Zealand Inc.,  
700 pp. 1363-1369.
- 701 Lombardi, F., Marchetti, M., Corona, P., Merlini, P., Chirici, G., Tognetti, R., Burrascano, S.,  
702 Alivernini, A., Puletti, N., 2015. Quantifying the effect of sampling plot size on the  
703 estimation of structural indicators in old-growth forest stands. *For. Ecol. Manage.* 346, 89-97.
- 704 Lundqvist, L., 1994. Growth and Competition in Partially Cut Sub-Alpine Norway Spruce  
705 Forests in Northern Sweden. *For. Ecol. Manage.* 65, 115-122.
- 706 Magnussen, S., Mandallaz, D., Lanz, A., Ginzler, C., Næsset, E., Gobakken, T., 2016. Scale  
707 effects in survey estimates of proportions and quantiles of per unit area attributes. *For. Ecol.*  
708 *Manage.* 364, 122-129.
- 709 Maltamo, M., Eerikäinen, K., Packalén, P., Hyyppä, J., 2006. Estimation of stem volume  
710 using laser scanning-based canopy height metrics. *Forestry* 79, 217-229.
- 711 Mascaro, J., Detto, M., Asner, G.P., Muller-Landau, H.C., 2011. Evaluating uncertainty in  
712 mapping forest carbon with airborne LiDAR. *Remote Sens. Environ.* 115, 3770-3774.
- 713 Matos, A., 2014. Effect of scale factor in estimation of Gini coefficient. Master's thesis.  
714 University of Eastern Finland.  
715 [http://epublications.uef.fi/pub/urn\\_nbn\\_fi\\_uef-20140718/index\\_en.html](http://epublications.uef.fi/pub/urn_nbn_fi_uef-20140718/index_en.html)  
716 [http://www.oppi.uef.fi/opk/video/europeanforestry/a\\_matos\\_seminar.mp4](http://www.oppi.uef.fi/opk/video/europeanforestry/a_matos_seminar.mp4)  
717 (accessed August 2016)
- 718 Mauro, F., Valbuena, R., Manzanera, J., García-Abril, A., 2010. Influence of Global  
719 Navigation Satellite System errors in positioning inventory plots for tree-height distribution  
720 studies *Canadian journal of forest research* 41, 11-23.
- 721 Mauro F., Molina I., García-Abril A., Valbuena R. & Ayuga-Téllez E. (2016) Remote  
722 Sensing Estimates and Measures of Uncertainty for Forest Variables at Different Aggregation  
723 Levels. *Environmetrics* 27(4): 225-238

- 724 McElhinny, C., Gibbons, P., Brack, C., Bauhus, J., 2005. Forest and woodland stand  
725 structural complexity: Its definition and measurement. *For. Ecol. Manage.* 218, 1-24.
- 726 McGaughey, R., 2015. FUSION/LDV: Software for LIDAR data analysis and visualization.  
727 Version 3.50. Forest Service. Pacific Northwest Research Station. United States Department  
728 of Agriculture. [Accessed December, 2015].
- 729 McRoberts, R.E., Winter, S., Chirici, G., LaPoint, E., 2012. Assessing Forest Naturalness.  
730 *For. Sci.* 58, 294-309.
- 731 Motz, K., Sterba, H., Pommerening, A., 2010. Sampling measures of tree diversity. *For. Ecol.*  
732 *Manage.* 260, 1985-1996.
- 733 Næsset, E., 2002. Predicting forest stand characteristics with airborne scanning laser using a  
734 practical two-stage procedure and field data. *Remote Sens. Environ.* 80, 88-99.
- 735 Næsset, E., 2004. Accuracy of forest inventory using airborne laser scanning: evaluating the  
736 first Nordic full-scale operational project. *Scand. J. For. Res.* 19, 554-557.
- 737 Næsset, E., 2007. Airborne laser scanning as a method in operational forest inventory: Status  
738 of accuracy assessments accomplished in Scandinavia. *Scand. J. For. Res.* 22, 433-442.
- 739 Næsset, E., Bollandsås, O.M., Gobakken, T., Solberg, S., McRoberts, R.E., 2015. The effects  
740 of field plot size on model-assisted estimation of aboveground biomass change using  
741 multitemporal interferometric SAR and airborne laser scanning data. *Remote Sens. Environ.*  
742 168, 252-264.
- 743 Neumann, M., Starlinger, F., 2001. The significance of different indices for stand structure  
744 and diversity in forests. *For. Ecol. Manage.* 145, 91-106.
- 745 O'Hara, K.L., Gersonde, R.F., 2004. Stocking control concepts in uneven-aged silviculture.  
746 *Forestry* 77, 131-143.
- 747 O'Hara, K.L., Hasenauer, H., Kindermann, G., 2007. Sustainability in multi-aged stands: an  
748 analysis of long-term plenter systems. *Forestry* 80, 163-181.
- 749 Oliver, C.D., Larson, B.C., 1990. *Forest Stand Dynamics*. McGraw-Hill, Inc.
- 750 Otypková, Z., Chytrý, M., 2006. Effects of plot size on the ordination of vegetation samples.  
751 *Journal of Vegetation Science* 17, 465-472.
- 752 Packalen, P., Vauhkonen, J., Kallio, E., Peuhkurinen, J., Pitkänen, J., Pippuri, I., Strunk, J.,  
753 Maltamo, M., 2013. Predicting the spatial pattern of trees by airborne laser scanning. *Int. J.*  
754 *Remote Sens.* 34, 5154-5165.
- 755 Pike, R.J., Wilson, S.E., 1971. Elevation-relief ratio, hypsometric integral, and geomorphic  
756 area-altitude analysis. *Geological Society of America Bulletin* 82, 1079-1084.
- 757 Pommerening, A., 2002. Approaches to quantifying forest structures. *Forestry* 75, 305-324.

- 758 Pommerening, A., Stoyan, D., 2006. Edge-correction needs in estimating indices of spatial  
759 forest structure. *Canadian Journal of Forest Research* 36, 1723-1739.
- 760 Pukkala, T., Laiho, O., Lähde, E., 2016. Continuous cover management reduces wind  
761 damage. *For. Ecol. Manage.* 372, 120-127.
- 762 R Development Core Team, 2016. R: A language and environment for statistical computing.  
763 R Foundation for Statistical Computing, Vienna, Austria. ISBN 3-900051-07-0, URL  
764 <http://www.R-project.org>.
- 765 Robles, A., Rodríguez, M.A., Alvarez-Taboada, F. 2016. Characterization of wildland-urban  
766 interfaces using LiDAR data to estimate the risk of wildfire damage. *Revista de*  
767 *Teledeteccion*, 45: 57-69.
- 768 Ruiz, L.A., Hermosilla, T., Mauro, F., Godino, M. 2014. Analysis of the influence of plot size  
769 and LiDAR density on forest structure attribute estimates. *Forests*, 5 (5): 936-951
- 770 Singh, K.K., Chen, G., McCarter, J.B., Meentemeyer, R.K. 2015. Effects of LiDAR point  
771 density and landscape context on estimates of urban forest biomass. *ISPRS Journal of*  
772 *Photogrammetry and Remote Sensing*, 101: 310-322
- 773 Smith, H.F., 1938. An empirical law describing heterogeneity in the yields of agricultural  
774 crops. *The Journal of Agricultural Science* 28, 1-23.
- 775 Staudhammer, C.L., LeMay, V.M., 2001. Introduction and evaluation of possible indices of  
776 stand structural diversity. *Canadian journal of forest research* 31, 1105-1115.
- 777 Sterba, H., Ledermann, T., 2006. Inventory and modelling for forests in transition from even-  
778 aged to uneven-aged management. *For. Ecol. Manage.* 224, 278-285.
- 779 Upton, G., Fingleton, B., 1985. *Spatial Data Analysis by Example. Volume 1: Point Pattern*  
780 *and Quantitative Data*. John Wiley & Sons Ltd.
- 781 Valbuena, R., Packalén, P., Martín-Fernández, S., Maltamo, M., 2012. Diversity and  
782 equitability ordering profiles applied to study forest structure. *For. Ecol. Manage.* 276, 185-  
783 195.
- 784 Valbuena, R., Packalen, P., Mehtätalo, L., García-Abril, A., Maltamo, M., 2013a.  
785 Characterizing forest structural types and shelterwood dynamics from Lorenz-based  
786 indicators predicted by airborne laser scanning. *Canadian Journal of Forest Research* 43,  
787 1063-1074.
- 788 Valbuena, R., Maltamo, M., Martín-Fernández, S., Packalen, P., Pascual, C. and Nabuurs, G.,  
789 2013b. Patterns of covariance between airborne laser scanning metrics and Lorenz curve  
790 descriptors of tree size inequality. *Canadian Journal of Remote Sensing*, 39, pp. S18-S31.
- 791 Valbuena, R., Vauhkonen, J., Packalen, P., Pitkänen, J., Maltamo, M., 2014. Comparison of  
792 airborne laser scanning methods for estimating forest structure indicators based on Lorenz  
793 curves. *ISPRS Journal of Photogrammetry and Remote Sensing* 95, 23-33.



- 794 Valbuena, R., Eerikainen, K., Packalen, P., Maltamo, M., 2016. Gini coefficient predictions  
795 from airborne lidar remote sensing display the effect of management intensity on forest  
796 structure. *Ecol. Ind.* 60, 574-585.
- 797 Valbuena, R., Maltamo, M., Mehtätalo, L., Packalen, P., 2017. Key structural features of  
798 boreal forests may be detected directly using 1-moments from airborne lidar data. *Remote  
799 Sensing of Environment* 194: 437-446.
- 800 Vihervaara P., Mononen L., Auvinen A.P., Virkkala R., Lü Y., Pippuri I., Packalen P.,  
801 Valbuena R., Valkama J., 2015. How to Integrate Remotely Sensed Data and Biodiversity for  
802 Ecosystem Assessments at Landscape Scale. *Landscape Ecology* 30 (3): 501-516.
- 803 Watt, M.S., Meredith, A., Watt, P., Gunn, A., 2014. The influence of LiDAR pulse density on  
804 the precision of inventory metrics in young unthinned Douglas-fir stands during initial and  
805 subsequent LiDAR acquisitions. *New Zealand Journal of Forestry Science*, 44 (1), 9 p.
- 806 Weiner, J., 1990. Asymmetric competition in plant populations. *Trends in Ecology &  
807 Evolution*, 5(11), pp. 360-364.
- 808 Weiner, J., Solbrig, O.T., 1984. The meaning and measurement of size hierarchies in plant  
809 populations. *Oecologia* 61, 334-336.
- 810 Weiner J., Thomas S., 1986. Size variability and competition in plant monocultures. *Oikos*  
811 47: 211–222.
- 812 Whittaker, R.H., 1972. Evolution and measurement of species diversity. *Taxon* , 213-251.
- 813 Zheng, X., Xia, T., Yang, X., Yuan, T., Hu, Y., 2013. The Land Gini Coefficient and Its  
814 Application for Land Use Structure Analysis in China. *PLoS ONE* 8, 1-10.

815

816 **Table Titles**

817 **Table 1.** Properties of the study area.

818

819 **Table 2.** Summary of ALS metrics computed with FUSION and used in this research  
820 (McGaughey, 2015).

821

822 **Table 3.** For each radii, proportion of the total number of original field plots within the  
823  $\overline{GC}_{diff} < 0.05$  limit (*Criterion D*), and average number of trees contained within those plots.

824

825 **Table 4.** Maximum absolute correlation between field  $GC$  and ALS predictors (*Criterion II*).

826 See **Table 2** for description of ALS metrics.

827

828

### 829 **Figure Captions**

830

831 **Figure 1.** Reproduction of tree positions (dots) within an original field plot (red rectangle)  
832 surrounded by edge correction i.e. translation method (i.e. periodic boundary), and a sample  
833 of 10 random realizations of simulated concentric circular plots with radii sizing 1-3 m (for  
834 simplicity). Axes show both absolute (above) and relative (below) coordinates (respectively  
835  $X_{abs}, Y_{abs}$  and  $X_{rel}, Y_{rel}$  in **Eq. 2**).

836

837 **Figure 2.** Results of sensitivity analysis to select minimum numbers of simulations.  
838 Evolution for increasing radii of (a) mean  $\widehat{GC}$  values and (b) their standard errors for  $k =$   
839 10-2000 simulations.

840

841 **Figure 3.** *Criterion I.* Asymptotic representation showing the evolution of  $\overline{GC}_{diff}$  (at each of  
842 the 79 original field plots) for increasing (a) plot sizes  $s = 1-15$  m radius (corresponding area  
843 also shown in upper axis) (b) and sample size  $n = 1-50$  number of trees (shortened to  
844 enhance visualization).

845

846 **Figure 4.** *Criterion II.* Absolute of correlation  $|r|$  between  $GC$  values and selected ALS  
847 predictors (see legend, and explanations of ALS metrics in **Table 1** and section 2.4) for  
848 increasing (a) plot size  $s = 1-15$  m radius (corresponding area also shown in upper axis) (b)  
849 and sample size  $n = 1-90$  number of trees.

850

851 **Figure 5.** Minimum number of trees (sample size) to reach  $GC$  stabilisation in relation to the  
852 reference  $GC$  value obtained from the original field plot ( $GC_{ref}$ ).

853

854 **Figure 6.** Changes due to varying ALS scan densities in the absolute of correlation  $|r|$   
855 between  $GC$  values and ALS predictors. See explanations of ALS metrics in **Table 1** (section  
856 2.4).

857

#### 858 **Supplementary Materials**

859 **Supplementary Figure 1.** Histograms showing the distribution of the response variable –  
860  $\overline{GC}$  (vertical bars) – and the predictor variables – *Skewness*, *Cover*, *CRR*, *P99*, *StdDev*, *P50*  
861 and *P25* (horizontal bars) –. The resulting scatterplots between each response-predictor pair  
862 are also shown. For simplicity, only results for the optimal plot radius  $s^* = 9$  m are shown.

863

**Table 1.** Properties of the study area.

Parameter	Minimum	Mean	Maximum	SD
$N$ (stems·ha <sup>-1</sup> )	467	1298	3025	594
$G$ (m <sup>2</sup> ·ha <sup>-1</sup> )	14	25	44	7
$QMD$ (cm)	10	17	29	4

$N$ : stand density;  $G$ : basal area;  $QMD$ : quadratic mean diameter; SD: standard deviation.

Draft

**Table 2.** Summary of ALS metrics computed with FUSION and used in this research (McGaughey, 2015).

Symbol	Description	Forest Characteristics
<i>P50</i>	Median (i.e. 50 <sup>th</sup> percentile)	Average tree height
<i>StdDev</i>	Standard deviation	Variation in tree heights
<i>Skew</i>	Skewness	Tree dominance
<i>P25</i>	1 <sup>st</sup> quartile (i.e. 25 <sup>th</sup> percentile)	Presence of understorey
<i>P99</i>	99 <sup>th</sup> percentile	Dominant height
<i>CRR</i>	Canopy relief ratio = $(\text{Mean} - \text{Min}) / (\text{Max} - \text{Min})$	Vertical structure
<i>Cover</i>	Percentage of all returns above 2 m	Canopy cover

Draft

**Table 3.** For each radii, proportion of the total number of original field plots within the  $\overline{GC}_{diff} < 0.05$  limit (*Criterion I*), and average number of trees contained within those plots.

Plot radius (m)	Ratio of original field plots reaching stabilization (%)	Average sample size of trees based on simulations
1	25.3	1.1
2	41.1	2.0
3	70.8	3.7
4	94.9	6.5
5	91.4	10.2
6	100	14.6
7	100	19.9
8	100	26.1
9	100	33.0
10	100	40.7
11	100	49.3
12	100	58.7
13	100	68.9
14	100	79.9
15	100	91.7

**Table 4.** Maximum absolute correlation between field *GC* and ALS predictors (*Criterion II*). See

**Table 2** for description of ALS metrics.

ALS metric	Maximum correlation $\max r $	Optimal plot radius ( $s^*$ ; $m$ )	Optimal number of trees ( $n^*$ ; $m$ )
<i>Skew</i>	0.58	10	41
<i>Cover</i>	0.45	12	59
<i>CRR</i>	0.42	9	33

Draft

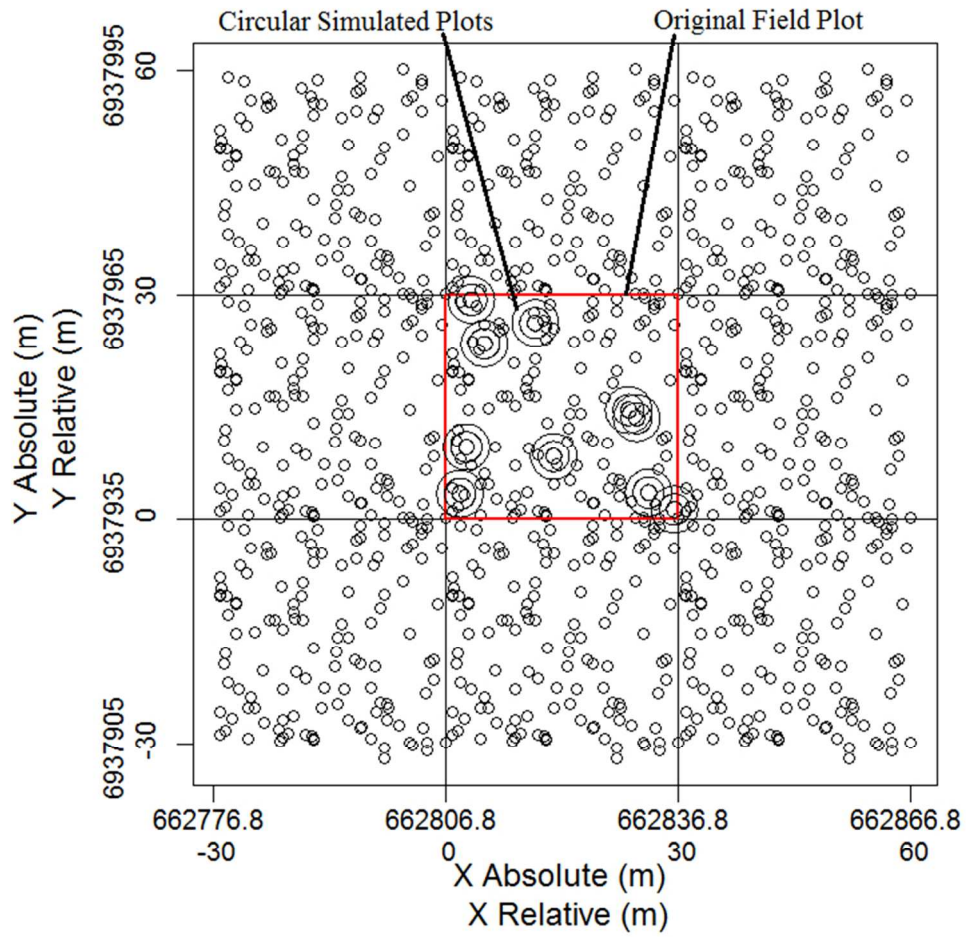


Figure 1

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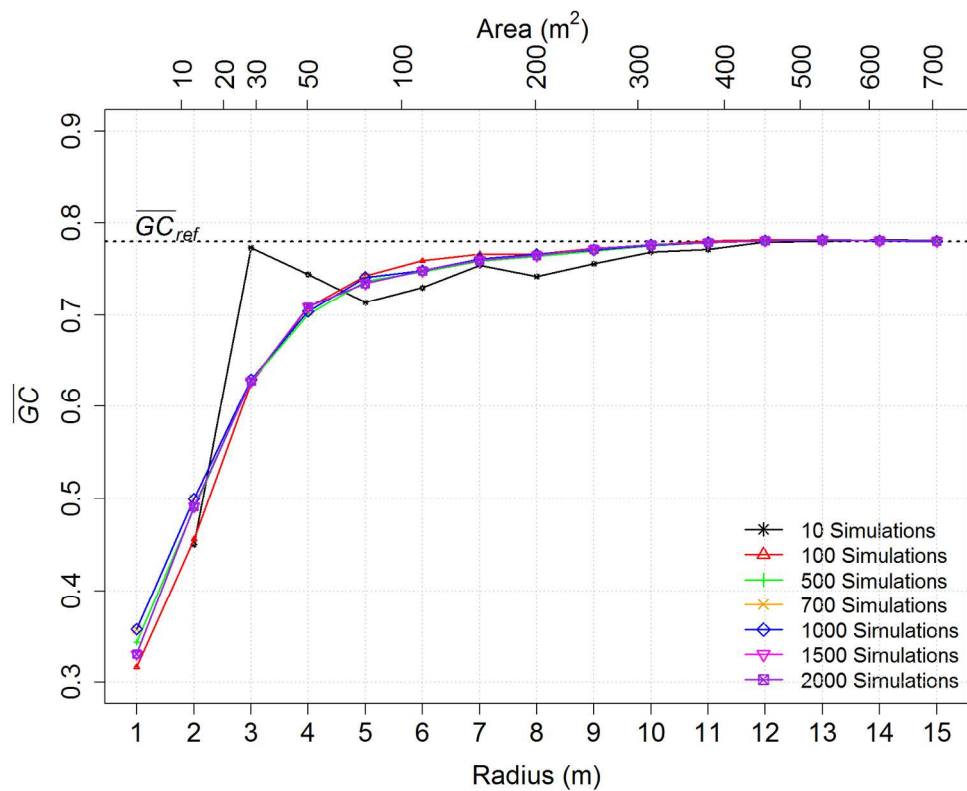


Figure 2(a)

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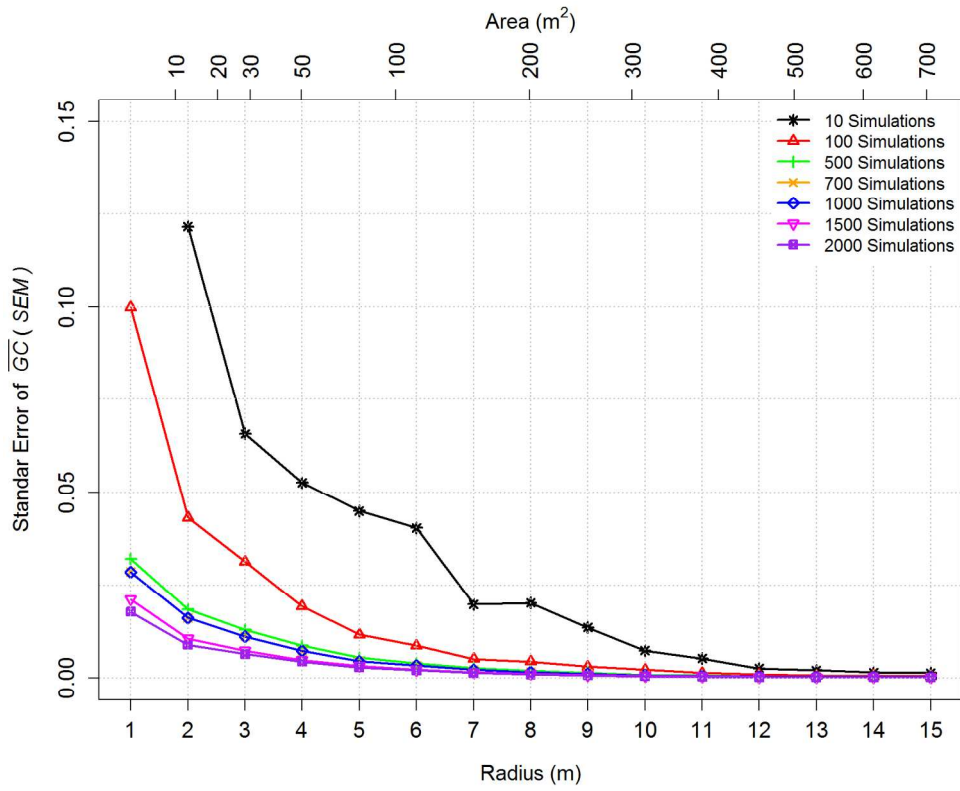


Figure 2(b)

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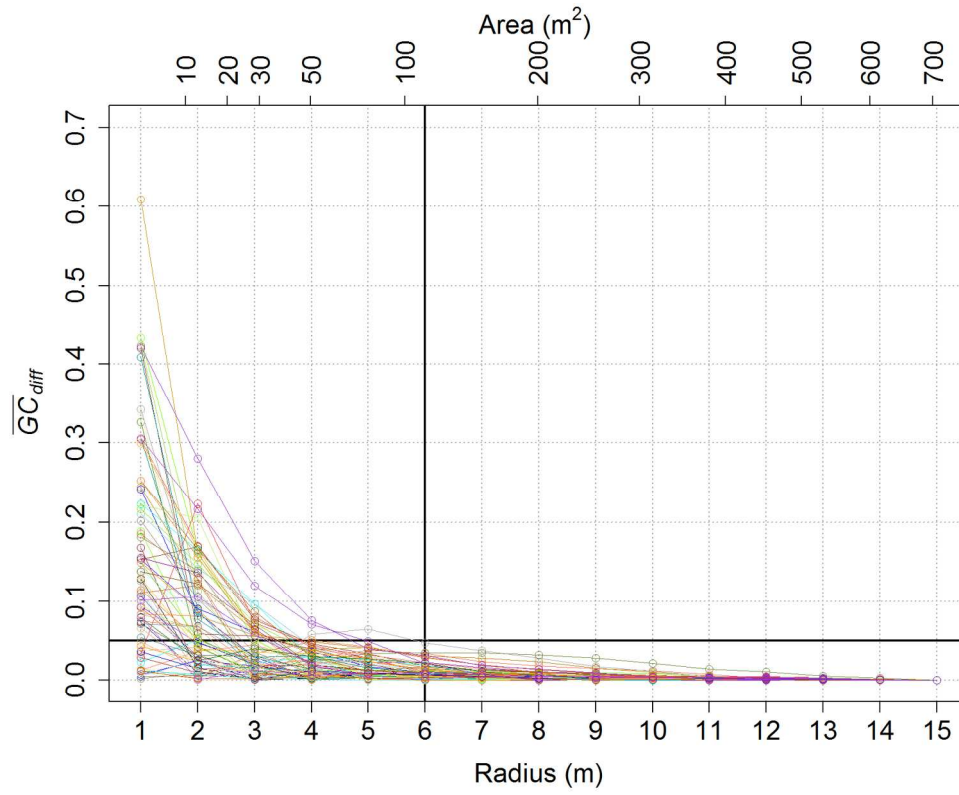


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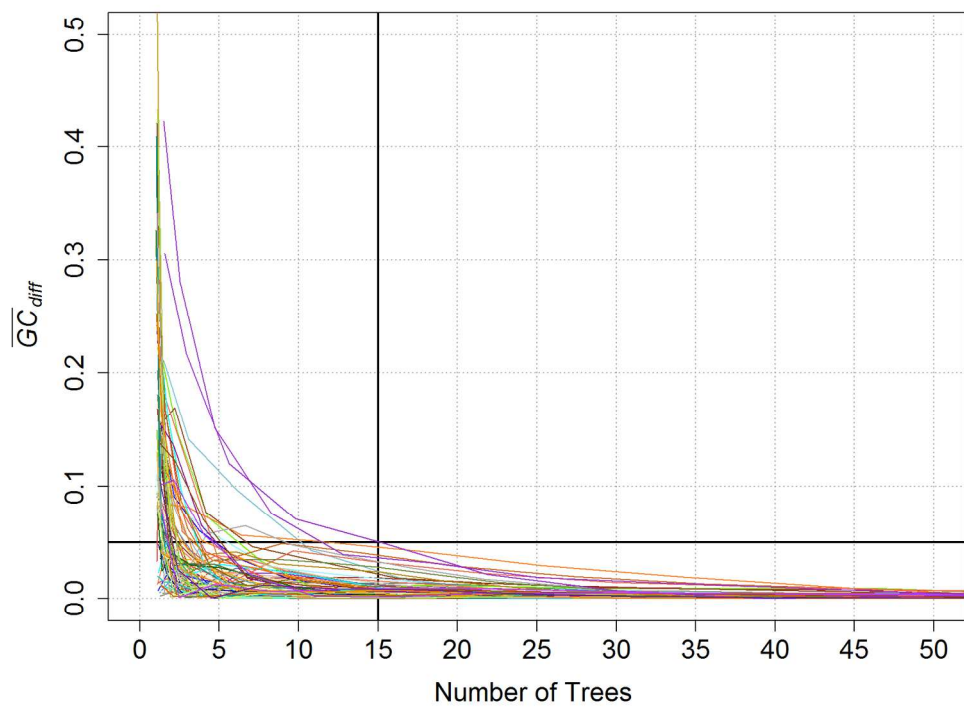


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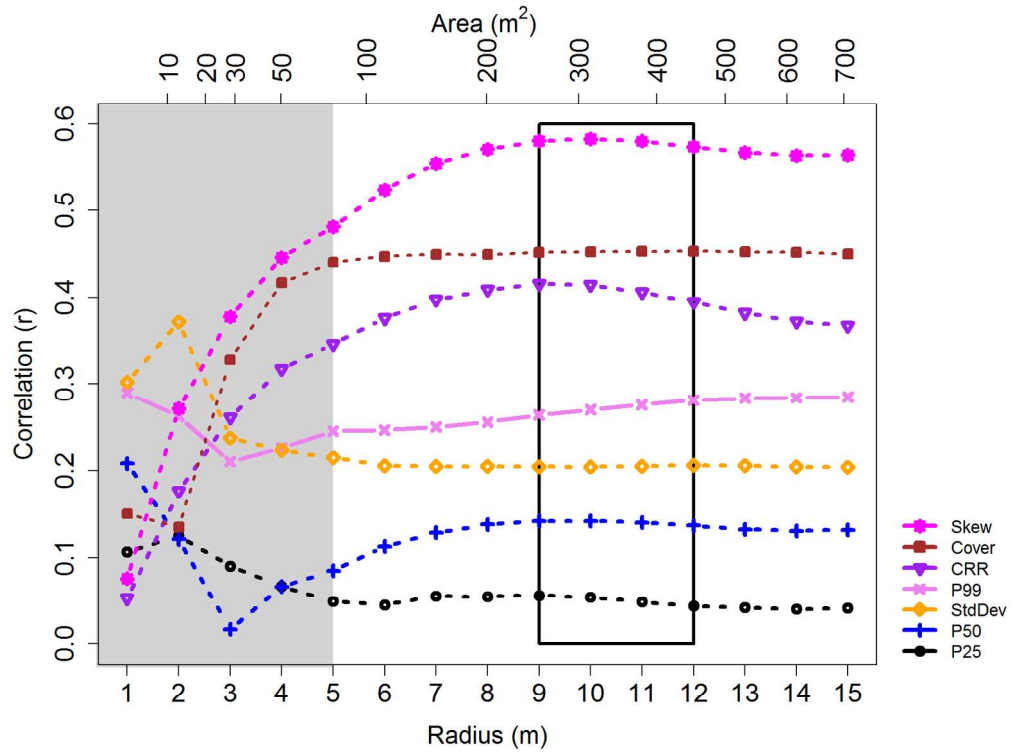


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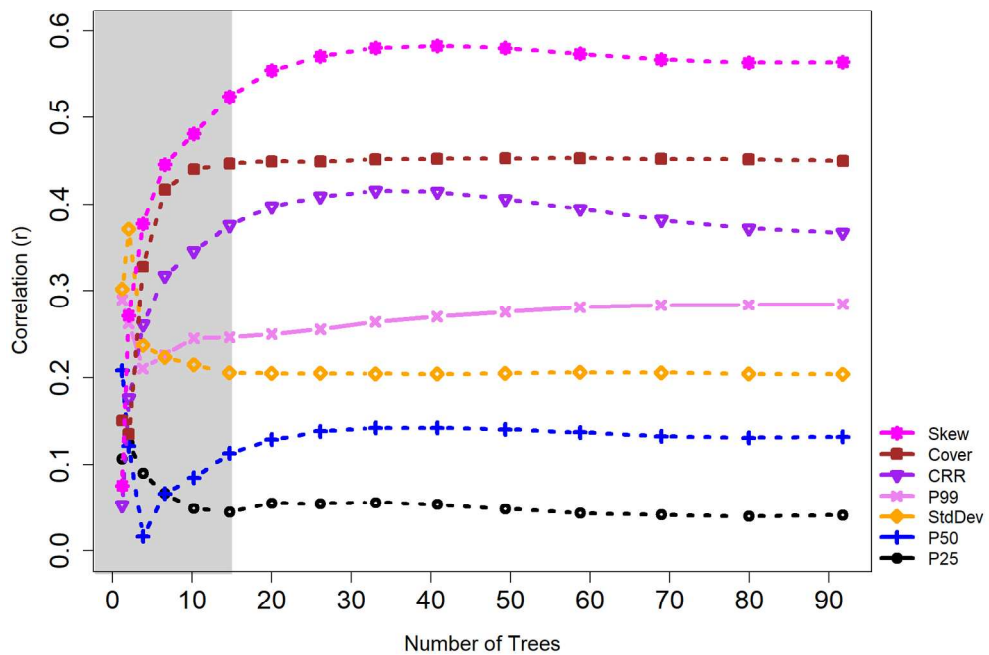


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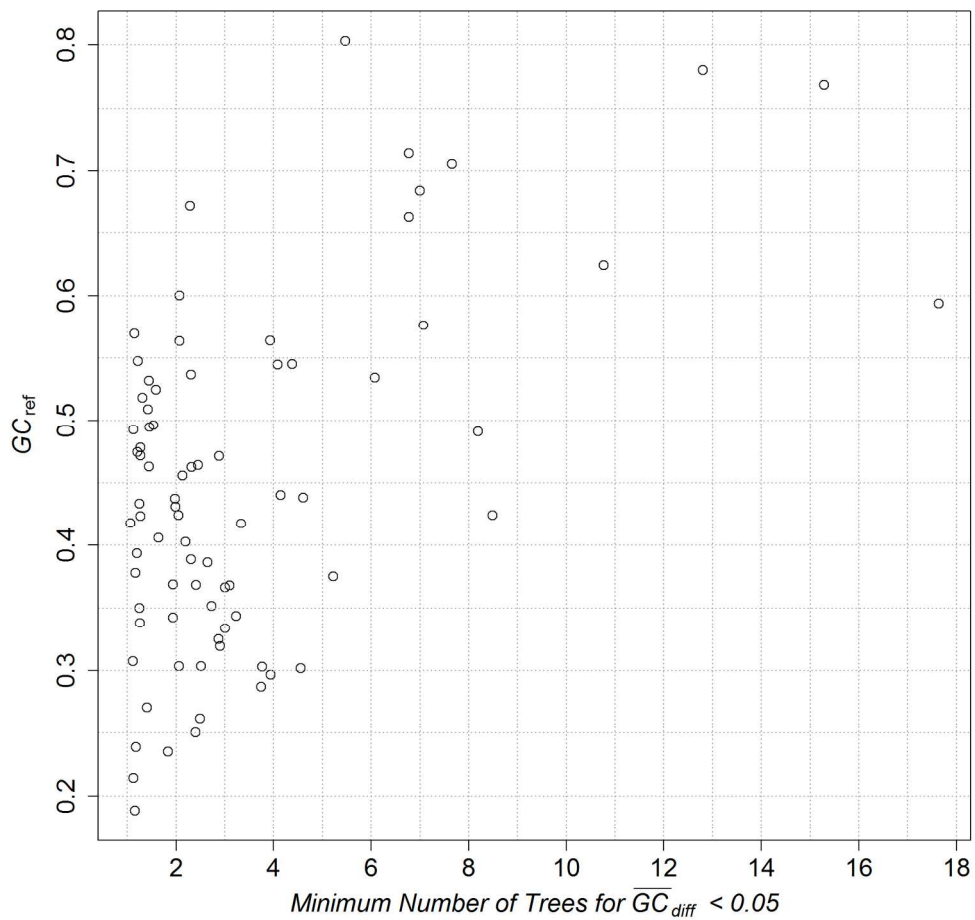


Figure 5

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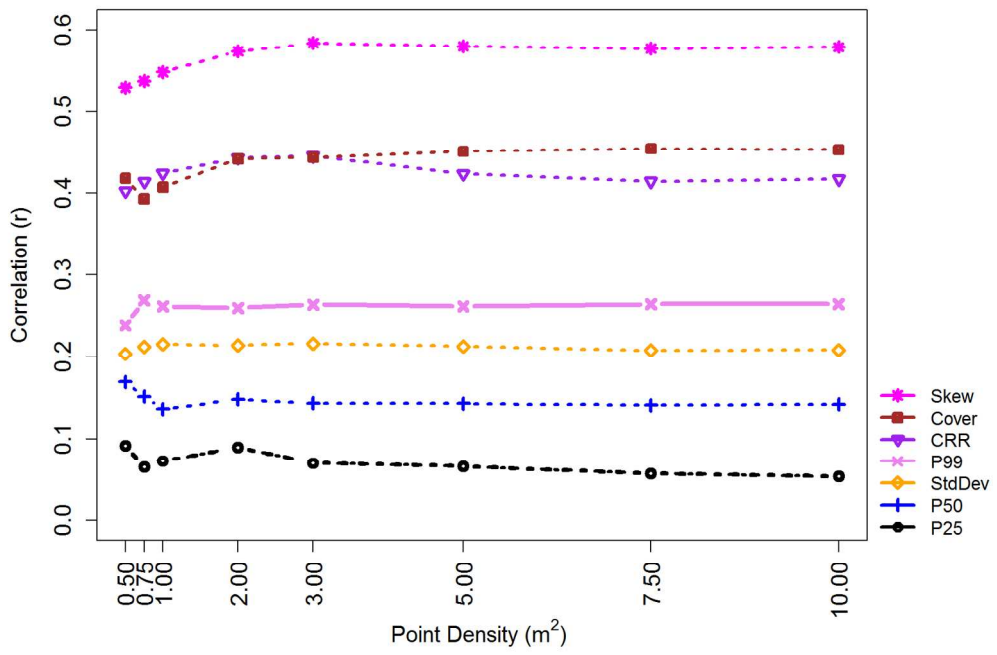
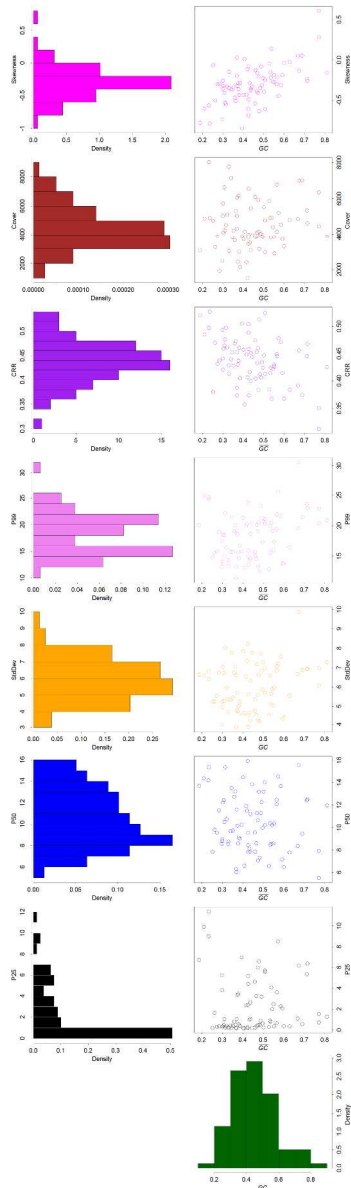


Figure 6

228x165mm (200 x 200 DPI)





558x1625mm (100 x 100 DPI)